APPENDIX G

Geologic and Hydrogeologic Information



MEMO



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December 10, 2019	30016349	
Subject:		
Groundwater Flow Model Development Waycross Yard, Waycross, Georgia	CSX Transportation, INC.,	

MODEL OBJECTIVES

In order to further understand groundwater flow and potential impacts of remedy pumping operations at the CSX Transportation, Inc (CSXT). Waycross Yard located in Waycross, Georgia (Site) a numerical groundwater flow model was developed. The groundwater flow model was developed using available Site and regional data and was calibrated to recent observed water levels. The calibrated groundwater flow model was then utilized to evaluate the hydraulic control of the delineated groundwater impacts under current remedy horizontal and vertical pumping wells.

GROUNDWATER FLOW MODEL

Groundwater Flow Modeling Software

The groundwater flow component of the modeling task was performed using the Modular Three-Dimensional Finite-Difference Groundwater Flow Model, widely known as MODFLOW (McDonald and Harbaugh, 1988). MODFLOW was originally distributed by the USGS. The version of MODFLOW used in this modeling task was based on the 1988 version MODFLOW, updated to include the most recent packages from the USGS. Key packages included with the current version of MODFLOW are an integration of the BCF3 (Goode and Appel, 1993) and LPF (Harbaugh et al., 2000) packages, spatially varying anisotropy (Kladias and Ruskauff, 1997), evaporation 2 package (Niswonger, et al., 2011), parameter estimation, and updated solvers (Niswonger, et al., 2011; Mehl and Hill, 2001). Data input is formatted, similar to MODFLOW-88 and MODFLOW-SURFACT (HydroGeoLogic, 1996), to enhance portability to other versions of MODFLOW and to simplify QAQC of data input files.

Model Domain

The groundwater flow model domain covers an area of approximately 3.9 square miles. The model domain extents were located a sufficient distance from the area of concern to reduce the potential for boundary effects. The northwestern and southern model boundaries align with unnamed intermittent surface water features. The northeastern model boundary aligns with the City Drainage Canal. The northern and eastern model boundaries align with interpolated regional groundwater flow paths and isocontours, respectively. The model was not rotated as the finite difference grid is generally aligned with the primary direction of groundwater flow in the vicinity of the Site which minimizes the potential for numerical erroring of the groundwater flow model. The groundwater flow model domain is shown in **Figure 1.**

Model Structure

The model was vertically discretized into 5 layers to represent the hydrostratigraphic units (HSUs) present at the Site as shown in cross-sectional view in **Figure 2**. The three more transmissive units comprised of silty sands and loose silty sands, Zones I, III, and V, are represented by model layers 1, 3, and 5, respectively. The less transmissive units comprised primarily of silty clay or sandy clay, Zones II and IV, are represented by model layers 2 and 4, respectively. The base of the model represents the stiff green clay of the Hawthorn Formation, as minimal groundwater flow occurs between the Hawthorn Formation and the overlying sediments. Model layer elevations were interpolated based on available Site boring logs, and changes in hydraulic conductivity zones were applied in areas where specific units were not encountered.

The finite difference grid is a telescopic mesh with finer grid cell resolution in the area of concern (Site) that grades up to a coarser grid cell dimensions towards the model domain extents. The grid cell dimensions vary from 10 ft x 10 ft in the area of concern up to 200 ft x 200 ft at the model perimeter (**Figure 1**). Adjacent grid cell dimensions do not vary by greater than 50% to reduce the potential for numerical error. The finite difference grid consists of 376 rows, 506 columns, and 5 layers for a total of 951,280 grid cells.

Hydraulic Parameters and Boundary Conditions

Recharge

Typical recharge from precipitation values are approximately 5% to 25% of the annual average precipitation rates. The average annual precipitation for the city of Waycross, Georgia is 50 in/yr (<u>https://www.usclimatedata.com/climate/waycross/georgia/united-states/usga0610</u>). The simulated recharge value refined during the model calibration was at the lower end of the typical recharge range at a value of 2.5 in/yr (5% of the annual precipitation rate).

Hydraulic Conductivity

Initial estimates for hydraulic conductivity were guided by available Site aquifer testing data and lithologic descriptions of the 5 zones encountered at the Site. During the model calibration process, the hydraulic conductivities were refined based on professional judgement. The final calibrated hydraulic conductivity values simulated in the model are shown in **Table 1**:

Model Layer	Corresponding HSU Zone	Horizontal Hydraulic Conductivity Range (ft/day)	Vertical / Horizontal Hydraulic Conductivity Ratio	Notes
1	Zone I	5.0	1 / 10	Limited monitoring well data available in Zone 1
2	Zone II	2.0	1 / 10	Higher than anticipated values supported by migration of mass through Zone II to Zone III
3	Zone III	5.0 to 15.0	1 / 10	Primary Zone of impacted groundwater
4	Zone IV	0.01	1 / 100	Holes in clay unit simulated where not present in boring logs; balance of unit relatively tight given the lack of response in Zone III when pumping HWW-5 in Zone V
5	Zone 5	3.0 to 7.0	1 / 10	Primarily characterized due to water level response near Zone V

Table 1. Simulated Hydraulic Conductivity Values

Surface Water Features

The regional unnamed intermittent surface water features were simulated as drain cell boundary conditions that only allow removal of water from the model if groundwater levels exceed elevation of a drain feature. The City Drainage Canals were simulated as river cells to either serve as a drain to groundwater, or a recharge source depending on the groundwater elevation relative to the defined elevation of the canal. Elevations of both the unnamed intermittent surface water features and City Drainage Canals were estimated from topographic maps of the area. The simulated surface water features are shown in **Figure 3**.

Constant Head Boundaries

The constant head boundaries were simulated at the extents of the active model domain to represent the regional groundwater flow through the deeper Zones of the model. The elevations of these constant head cells were interpolated from the regional groundwater flow through the Site. The constant head cells were located a sufficient distance from the Site to minimize the potential for boundary effects. The simulated constant head cells are shown in **Figure 3**.

Extraction Wells

The only active extraction wells within the model are associated with the Site recovery system consisting of 5 horizontal wells and 11 vertical wells. The cumulative rate for the vertical wells was evenly distributed across all 11 vertical wells as individual vertical well flow rates are not available. The horizontal wells were simulated as a continuous series of wells located only along the screened portion of the horizontal extraction well. The horizontal well extraction rate for each of the horizontal wells were then evenly distributed across the cells representing the screened portion of the horizontal well. The locations of the simulated extraction wells are shown in **Figure 3**.

Groundwater Flow Model Calibration

Calibration of a groundwater flow model refers to the iterative process of systematically adjusting the model boundary conditions and input parameters within a justifiable and generally accepted range of values to obtain as close a match between observed and simulated water levels. The generally-accepted process for comparison of simulated and observed water levels is summarized in ASTM Standard D-5490-93 (ASTM 1994) and utilizes the concept of residuals – which is the difference between the simulated and observed water level.

The residuals are evaluated using standard statistical measures; residual sum-of-squares (RSS), variance, and mean. The standard deviation (the square root of the variance) is the median error and needs to be small compared to the range in observed data. A well calibrated model will reproduce observed water levels with a maximum median error of 10% or less of the difference between the largest and smallest observed water level (Anderson and Woessner, 1992). The calibration process also evaluates the residuals using the residual mean.

The groundwater modelling calibration was performed under steady-state conditions based on data from a synoptic water-level round conducted on October 21, 2019 representative of groundwater flow conditions under active remedy pumping. The calibration data set consists of 128 observed Site water levels under the following pumping rates shown in **Table 2**:

Well	Туре	Rate (gpm)
HWW-1	Horizontal Well (Zone III)	34
HWW-2	Horizontal Well (Zone III)	10.3
HWW-3	Horizontal Well (Zone III)	5
HWW-4	Horizontal Well (Zone III)	6.4
HWW-5	Horizontal Well (Zone V)	6.5
Vertical Wells	11 Vertical wells	3.1

Table 2. Calibration extraction well rates

The resultant calibration is shown graphically in **Figure 4**. The scaled standard deviation is 10%, the residual mean is 0.1, and the scaled residual mean is 0.01. There is still local variation in the calibration residuals given the complexity and the heterogeneity of the hydrogeology encountered at the Site, but the overall residual statistics indicate a well calibrated groundwater flow model. The groundwater flow model calibration was also evaluated qualitatively by comparing the simulated and observed general groundwater flow directions and gradients. This qualitative assessment indicates that the model aligns

with observed groundwater flow orientations. The groundwater flow model was run for both non-pumping conditions and recent active pumping rates as of November 2019 as summarized in **Table 3**.

Well	Туре	Rate (gpm)
HWW-1	Horizontal Well (Zone III)	28
HWW-2	Horizontal Well (Zone III)	10
HWW-3	Horizontal Well (Zone III)	5
HWW-4	Horizontal Well (Zone III)	6
HWW-5	Horizontal Well (Zone V)	7
Vertical Wells	11 Vertical wells	3

Table 3. Current extraction well rates

gpm = gallons per minute

The simulated water levels under non-pumping conditions for Zones III and V are shown in **Figures 5 and 6**, respectively.

Hydraulic Capture Assessment

Upon completion of the calibration of the groundwater flow model, hydraulic control of Zone III and V groundwater by the remedy wells was assessed by utilizing the volumetric flux model MODular flow ALLocation (MODALL) program (Potter et al., 2008). MODALL uses the same MODFLOW-calculated cell-by-cell flow output as MODPATH (Pollock, 1989), but rather than tracking individual particle paths, it tracks the movement of groundwater in the model. This approach, also described as the volumetric-tracking method, is described in detail by Barlow et al. (2018). For the purpose of this analysis, the recent active pumping rates from November 25, 2019 were evaluated as shown in **Table 3**.

The simulated hydraulic capture under current conditions for Zones III and V are shown in **Figures 7 and 8**, respectively. These simulated water levels are consistent with the observed (October 2019) water levels in Zone III and Zone V strata, shown in **Figures 9 and 10** respectively, in that cones of depression develop around the active horizontal and vertical extraction wells influencing the flow directions in both hydrostratigraphic zones. It must be noted that horizontal wells HWW-1 and HWW-5, both screened in Zone V strata, capture a portion of groundwater from Zone III due to the lack of Zone IV aquitard strata at and upgradient of the well screens. In addition, HWW-5 also captures Zone V groundwater over a significant area despite the relatively low operational extraction rate of 7 gpm. The remaining Zone III wells (HWW-2, HWW-3, HWW-4, and the vertical wells) have minimal hydraulic influence on Zone V groundwater due to the presence of the competent Zone IV aquitard throughout the majority of the Site.

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Table 1 Summary Hydraulic Conductivity-Acid-Lime Sludge Area SWMU **CSX** Transporation, Inc. Waycross, Georgia

ALSA Zone III Monitor Wells	Slug Out	
	cm/sec	ft/day
MW-6	0.003825	10.842536
MW-21	0.0005993	1.698806
MW-22	0.001488	4.217959
MW-56	0.001241	3.517801
MW-135S	0.001071	3.035910
Average Conductivity Zone III	0.00164486	4.662602

Average Conductivity Zone III

0.00164486

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ALSA Zone V Mo	nitor Wells		
MW-18D		0.0001299	0.368221
MW-21D		0.0002555	0.724253
MW-51D		0.002535	7.185838
MW-52D		0.005808	16.463647
MW-116D		0.001403	3.977014
MW-122D		0.0001176	0.333355
MW-134D		0.0009834	2.787595
MW-135D		0.002264	6.417648
MW-136D		0.001444	4.093235
MW-137D		0.002456	6.961900
	Average Conductivity Zone V	0.00173964	4.931270526

Average Conductivity Zone V

4.931270526

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DRAFT PARCADIS Design & Consultancy for natural and built assets

Notes:

ALSA=Acid Lime Sludge Area Solid Waste Management Unit cm/sec=centimeters per second ft/day=feet per day Coversion from cm/sec to ft/day is 1 cm/sec=2,834.65 ft/day



MW-6 SLUG OUT

Data Set: <u>G:\AProject\CSXT\WaycrossRCRA\SC000239.0044 ALSA GW\Slug Tests\MW-6 Slugout.aqt</u> Date: <u>12/12/19</u> Time: <u>15:55:07</u>

PROJECT INFORMATION

Company: <u>ARCADIS</u> Client: <u>CSX Waycross</u> Project: <u>SC000239.0044</u> Test Location: <u>Waycross, GA</u> Test Well: <u>MW-6</u> Test Date: <u>6/3/2016</u>

Test Date: $6/3/2016$			
AQUIFER DATA			
Saturated Thickness: 20.29 ft	Anisotropy Ratio (Kz/Kr): 1.		
W	ELL DATA (MW-6 Slugout)		
Initial Displacement: <u>2.6</u> ft Casing Radius: <u>0.166</u> ft Screen Length: <u>10.</u> ft	Water Column Height: 20.29 ft Wellbore Radius: 0.5 ft		
	SOLUTION		
Aquifer Model: Unconfined	Solution Method: Bouwer-Rice		

K = 0.003825 cm/sec

y0 = 1.638 ft






































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CSX Transportation, Inc. Old Drum Storage Area HWMU

Assumptions

- Anisotropy ratio (Kz/Kr) = 1
- Saturated Thickness = 19 feet based on pre-pumping water levels at WW-44, WW-45, WW-46, boring logs for those wells, and nearby HPT logs
- Overlying aquitard thickness (B') = 11 ft based on pre-pumping water levels at WW-44, WW-45, WW-46, boring logs for those wells, and nearby HPT logs
- Underlying aquitard thickness (B") = 1 ft. Note: Results are insensitive to this parameter.
- All pumping and observation wells are fully penetrating based on boring logs and HPT logs
- Leaky confined aquifer

Observation Well	Storativity	Transmissivity	Saturated Thickness	Hydraulic Conductivity	Hydraulic Conductivity	
(Well ID)	(dimensionless)	(ft2/min)	(ft)	(ft/day)	(cm/sec)	
MW-36	3.00E-03	7.3E-03	19	0.55	1.9E-04	
MW-72	1.15E-03	1.4E-02	19	1.03	3.6E-04	
MW-73	3.59E-05	3.7E-02	19	2.83	1.0E-03	
MW-74	9.42E-06	2.7E-02	19	2.07	7.3E-04	
MW-76	2.81E-05	3.2E-02	19	2.41	8.5E-04	
MW-77	1.04E-05	1.9E-02	19	1.46	5.1E-04	
Maximum	3.00E-03			2.83	1.0E-03	
Minimum	9.42E-06			0.55	1.9E-04	
Geometric Mean	8.36E-05			1.51	5.3E-04	













ASB PUMP TEST DATA AND ANALYSIS SEPTEMBER 16, 1987

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Aquifer Test Results Waycross, Georgia Facility

80H- received 19/12

Prepared for:

Georgia Environmental Protection Division Atlanta, Georgia

For:

CSX Transportation Jacksonville, Florida

October 1987

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AQUIFER TEST RESULTS WAYCROSS, GEORGIA FACILITY

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440-03-05 October, 1987

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Prepared For:

CSX Transportation Jacksonville, Florida

Prepared By:

ERT, A Resource Engineering Company 3000 Richmond Houston, Texas 77098

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1.0 OBJECTIVES

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On April 2, 1987, ERT performed a slug test at each of two hazardous waste management areas (HWMA) at CSX Transportation's Waycross, Georgia Facility. These slug tests were performed at the Old Drum Storage Area (ODSA) and the Alum Sludge Basin (ASB) to roughly define the hydraulic properties of the uppermost estimated an results, slug test the aquifer. From and average linear conductivity, hydraulic transmissivity, velocity of the aquifer were obtained for both units. Based on these slug test results the estimated hydraulic properties were calculated as follows:

Old Drum Storage Area (ODSA) Estimated Results

Transmissivity	T =	339 gpd/ft
Hydraulic Conductivity	K =	48.4 gpd/ft ²
Average Linear Velocity	V =	29.96 ft/year

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Alum Sludge Basin (ASB) Estimated Results

Transmissivity T = 641 gpd/ftHydraulic Conductivity $K = 64 \text{ gpd/ft}^2$ Average Linear Velocity V = 177.23 ft/year

These properties were used in designing an aquifer pumping test. The results of that pumping test would enable ERT to calculate predicted drawdowns at the two HWMA during corrective action withdrawal and treatment. These predicted drawdowns are necessary for the determination of withdrawal well spacing and pumping rates.

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2.0 WELL INSTALLATION

From August 8 to August 15, 1987, ERT conducted a pumping test well installation program at the Alum Sludge Basin, where the pumping test would be performed. A network of six pumping wells was installed as shown in Figure 2-1. Well test installation was carried out under direct supervision of ERT in accordance with Standard Operating field geologists and (SOP) presented in Appendix A of this report. This Procedures program was performed using a truck-mounted drill riq and standard geotechnical practice. All monitor well bore holes were drilled and advanced by the mud-rotary method using a mixture of The field geologist Aqua-Gel and city water as a drilling fluid. and recorded changes in lithology by a visual identified examination of soil samples and the drill cuttings. Soil samples were collected at five foot intervals and/or at changes in lithology using a standard split-barrel sampler in accordance with ASTM designation D-1586, "Penetration Test and Split Barrel Soil samples were visually examined, Sampling of Soils". logged by the field geologist according to the classified, and Soil Classification System. Boring logs were prepared Unified each well/piezometer from this information and are provided for in Appendix B of this report.

Each pumping test well receives groundwater through a Schedule 40 Tri-Loc threaded, polyvinyl chloride (PVC) screen placed in the wash boring drilled to a predetermined depth. Two wells in the pumping test network are six inches in diameter (PW-1 and OB-3); the remainder of the wells are two inches in diameter.

A sand filter was placed in the annulus around the screen. The filter was then sealed at the top with a bentonite seal and the annulus above was backfilled with a bentonite-cement grout

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mixture. All construction material and drilling/sampling equipment was decontaminated before each use in accordance with ERT's Standard Operating Procedures (SOP). The decontamination procedure included, in order: steam cleaning, isopropanol rinse, deionized water rinse and air drying.

A typical well design is shown in Figure 2-2. After completion, each well was developed using a hand pump to evacuate any drilling fluid and/or residual sediment. Construction details are summarized as follows in Table 2-1.

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TABLE 2-1 PUMP TEST WELL CONSTRUCTION DETAILS CSX TRANSPORTATION WAYCROSS, GEORGIA

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Well No.	Elevations + Top of Pipe	Casing Type	Length*	Screer I.D.	Slot size	Length (ft)	Depth (ft)	Filter Dep (ft)
	msl (ft)				7	12.1	· · · · · · · · · · · · · · · · · · ·	
1-11d	136.840	А	13.5	9	0.020	10	* 17-27	15-27
	136.410	A	14.0	5	0.010	10++	17-27	14.8-32
OB-2	136.850	А	14.0	2	0.010	10++	17-27	14-32
OD-E DR_3	137.210	A	12.5	9	0.020	10++	. 16-26	14-31
	136 690	A	14.0	2	0.010	10++	17-27	15-32
00-4	135 880	: A	13.0	2	0.010	10++	17-27	14.5-32
0-00 MJ-15	135.990	<u> </u>	18.0	7	0.010	10++	18-28	17-33
MW-31	136.410	в	17.0	2	0.010	10++	17-27	14-32
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A- 2-inch or 6-inch diameter Schedule 40 Tri-loc pvc pipe
B- 2-inch Type 304 Tri-loc stainless steel pipe
+- Elevations (msl) as reported by H: W. Williams & Assoc., Inc., Waycross, Georgia
*- Casing length from ground surface
++- 5-foot sediment trap attached at bottom of screen

3.0 AQUIFER TEST

3.1 Field Program

Two aquifer pumping tests were performed on September 16, 1987 using an In Situ model SE200 Hydrologic Analysis System. This system consists of a computer which receives water level data from several transducers (in this case, five) in observation wells while one well is being pumped at a constant rate to produce a drawdown. The data generated by the computer during the two tests are provided in Appendix C of this report. A diagram showing the piping used during the tests is provided in Figure 3-1. Discharge water was containerized using a tank truck and was disposed of, as per GAEPD approved procedure, at Wastewater Treatment Plant No. 3 on site.

Two tests were performed in order to confirm the data obtained. The first test was conducted with the submersible pump located near the top of the screen in the pumped well (PW-1). The second test was conducted with the submersible pump located near the bottom of the PW-1 screen. Each test consisted of a drawdown phase and a recovery phase (falling water levels were monitored while pumping, and rising water levels were monitored after pumping was terminated). Comparison between the two phases gives an idea of the degree of variation in the data.

3.2 Conclusions

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From the data obtained during the tests, transmissivity and storage coefficient are calculated. This was accomplished using Jacob's approximation method. This method involves the plotting of data as a curve and measuring desired parameters from the curve. The curves plotted for ERT's data are presented in

3-1





Appendix C of this report (note that drawdown data for Test 2 Input 1 cannot be plotted due to the interfering effect of water cascading over the screen in the pumped well). Table 3-1 provides measured transmissivities and storage coefficients from those curves.

Theoretically, all calculations of transmissivity for these two tests should produce the same value. Due to variations in the aquifer, however, and slight fluctuations in factors such as pumping rate and even weather conditions, values normally show some range of variability. Considering that the hydraulic conductivity of an aquifer (proportional to the transmissivity) can vary over a range spanning more than 13 orders of magnitude, the values obtained from the September 16 aquifer tests are relatively consistent.

order to predict drawdowns, in the withdrawal wells, data In selected from that portion of the tests which most closely was approximates the conditions which will exist in the withdrawal wells - that is, drawdown of water levels with the pump situated near the bottom of the screen. The values from Test 2, drawdown phase, represent such a situation, therefore those values were averaged to obtain a mean transmissivity of 1983 gpd/ft and a mean storage coefficient of 2.213 x 10^{-3} . These means were used to predict drawdowns for withdrawal wells at the Alum Sludge Basin as presented in Appendix D of this report. Based on ERT's initial slug tests, transmissivity at the Old Drum Storage Area just over half that at the Alum Sludge Basin, so a is transmissivity of 1,000 gpd/ft was assumed for the ODSA, and the storage coefficient, 2.213×10^{-3} , was used (see Appendix same These values can be used to calculate other aquifer D). characteristics as shown below:

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TABLE 3-1

AQUIFER CHARACTERISTICS CALCULATED FROM PUMPING TEST DATA

Test #1

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Pumping Rate Q = 1.8 gpm Jacob's Approximation Method

	Transmissiv	ity T, gpd/ft	<u>Storage</u> Coe	fficient S
Well No.	Drawdown	Recovery	Drawdown	Recovery
OB-1	1218	1131	9.75×10^{-4}	1.06×10^{-3}
OB-5	1533	1760	5.03×10^{-4}	5.58×10^{-4}
MW-31	2160	2376 🚦	3.11×10^{-3}	4.89×10^{-3}
MW-15	848	969	1.64×10^{-3}	1.27×10^{-3}

Test #2

Pumping Rate Q = 3.5 gpm Jacob's Approximation Method

	Transmissiv:	ity T, gpd/ft	<u>Storage</u> Coe	efficient S
Well No.	Drawdown	Recovery	Drawdown	Recovery
OB-1	1200	943	1.41×10 ⁻³	1.32x10 ⁻³
OB-3	3787	*	1.10×10^{-3}	*
MW-31	2174	3029	7.51×10^{-4}	7.5×10^{-4}
MW-15	770	616	1.77×10^{-3}	1.64x10 ⁻³

* Not enough response shown in well during test to obtain reliable data for this value.

Old Drum Storage Area

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Transmissivity	T = 1,000 gpd/ft
Hydraulic Conductivity	$K = 143 \text{ gpd/ft}^2$
Storage Coefficient	$S = 2.213 \times 10^{-3}$
Average Linear Velocity	V = 58 ft/year

Alum Sludge Basin

Transmissivity	Т	=	1,983 gpd/ft
Hydraulic Conductivity	K	=	198 gpd/ft ²
Storage Coefficient	S	=	2.213×10^{-3}
Average Linear Velocity	v	=	385 ft/year

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3-5



4.0 CHANGES TO PART B PERMIT APPLICATION

Sections 3.4.1.3 and 4.4.1.3 of ERT's "Revised Part B Post-Closure Permit Application - Waycross, Georgia Facility" dated August, 1987 provide conceptual designs for the withdrawal well systems at the ASB and OSDA. The values for transmissivity, storage coefficient, pumping rate, and drawdowns given in the above mentioned sections are initial estimates and are now superseded by the values determined during the September 16 pumping tests.

The new-values, however, still support the design presented in the Part B Permit Application; the number of wells, construction details, well spacings, locations, and withdrawal procedures remain as originally planned:

LSA PUMP TEST DATA AND ANALYSIS APRIL 19, 1994 THROUGH APRIL 21, 1994

LOCOMOTIVE SKOP AREA), NAY CROSS, GEORGIA - JUNE 1995 eder associates

- three sets of 2-inch diameter, 0.02-inch slot, PVC wells to monitor different depth zones during the pilot test (MP-1S, MP-1M, MP-1D, MP-2S, MP-2D, MP-3S, and MP-3D); and
- one 2-inch diameter, 0.02-inch slot, PVC air sparging well screened in Zone III to be used if sufficient drawdown was obtained to allow soil vapor recovery (AS-1).

Continuous split-spoon samples were collected at each boring location (except AS-1) to determine soil lithology. Well boring logs are included in Appendix D. The cross sections on Figure 12 show the shallow lithology and the well screen intervals.

These wells were installed in accordance with GAEPD's "Manual for Groundwater Monitoring." Well construction logs are included in Appendix D. Upon completion of well construction, each well was developed using surge-block and pumping/purging techniques until clear water was obtained. VG-2 was developed for approximately six hours to ensure well efficiency.

10.5.4 48-Hour Aquifer Test

The 48-hour aquifer test was done to determine aquifer hydraulic characteristics in the Locomotive Shop Area. (April 19,1994 + though April 21,1994)

Equipment used during the aquifer test included:

- an In-Situ eight channel data logger and depth transducer system for collecting groundwater level data;
- a Grundfos submersible pump; and

• a 6,000-gallon tanker for groundwater discharge storage and transport to the on-site air stripper system.

Before the 48-hour aquifer test, the data logger system was used to evaluate antecedent groundwater level fluctuations. No fluctuations were detected. Antecedent groundwater levels are included in Appendix E.

A step drawdown test was performed to determine a sustainable yield rate for the 48-hour aquifer test. Pumping rates of 1.0, 2.0, 2.1, 2.3, 2.5, 2.7, and 3.0 gpm were evaluated. The flow rate was set at a given step and an In-Situ Data logger and depth transducer were used to measure the water level drawdown at the given flow rate. At the 1.0 and 2.0 gpm flow rates, the water levels stabilized at a higher level than the maximum drawdown available in the recovery well. At flow rates higher than 2.1 gpm, the water level did not stabilize and there was clear evidence that continued pumping at that rate would result in dry well conditions. At the 2.1 gpm flow rate, the water level stabilized just above the intake of the withdrawal pump, allowing continued pumping at maximum drawdown for an indefinite period of time. The 2.1 gpm sustainable flow rate was confirmed during the 48-hour aquifer test.

During the 48-hour aquifer test, groundwater samples were collected from VG-2 at the beginning, middle, and end of the 48-hour period. Laboratory data for these samples are included in Appendix E. TCE groundwater concentrations increased from 44,000 $\mu g/\ell$ to 72,000 $\mu g/\ell$ during the test.

Barometric pressure data was also collected during the 48-hour aquifer test to determine if any correction factors needed to be applied. No significant fluctuations in barometric pressure were observed. The data logger results were confirmed throughout the test using a manual electronic water meter. At the end of the 48-hour aquifer test, the pump was turned off and recovery data was collected until the groundwater returned to normal levels. This data is included in Appendix E.

The groundwater drawdown data indicated a very narrow cone of depression, with limited drawdown effect in monitoring points as close as 13 feet away. Detailed analysis of the aquifer test results are provided in Section 10.5.7.1.

10.5.5 Soil Vapor Extraction Test

The cone of depression of the water table at the recovery well was very narrow; consequently, insufficient drawdown was obtained to allow any significant or widespread SVE influence on Zone III soils. A limited SVE pilot test was conducted on Zone I soils by connecting a temporary blower to extraction well VG-2. VG-2 was pumped at 2.9 gpm to maintain drawdown. The blower used for the SVE test was a GAST Model R51250-50 with a 2.5 horsepower motor. The blower maintained a flow rate of approximately 100 standard cubic feet per minute (cfm) at a vacuum of 38 inches H₂0. The vacuum was monitored during system operation at monitoring points MP-1S, MP-2S, and MP-3S. The system was operated for 98 minutes to allow the vacuum to stabilize. Vacuum response at MP-1S (16 feet from VG-2) and MP-2S (40 feet from VG-2) stabilized at 0.50-inches H₂0 and 0.04-inches H₂0, respectively. MP-3S (82 feet from VG-2) showed no response to the vacuum.

Based on the narrow cone of depression observed during the aquifer pump test and the results of the SVE pilot test, soil vapor extraction is not effective technology for accelerating the remediation of Zone III or for remediating the limited soil contamination in Zone I at the former parts cleaning vat location. Since the soil contamination in Zone I is minimal and is capped by the concrete floor of the Locomotive Shop, there is no pathway for human exposure and no significant driving force for downward migration of the contaminants into the water table.

10.5.6 Air Sparging Test

The air sparging well (AS-1) was not used since drawdown was insufficient during the SVE test to allow capture of vapors in Zone III. Vapor bubbles in the saturated conditions of Zone III would not be able to migrate upward through the relatively impermeable soils of Zone II. These vapors would migrate laterally and could not be effectively captured. This could cause the plume to migrate outward from the air sparging point. Generally, if the vapors cannot be captured, air sparging should not be used.

10.5.7 Pilot Study Results

10.5.7.1 Analysis of Aquifer Test Results

Aquifer test results were analyzed to determine hydrogeologic conditions and establish aquifer parameters for transmissivity (T), storativity (S), and hydraulic conductivity (K). Aquifer test solution software (AqtesolvTM) was used to analyze the 48-hour drawdown data. All appropriate models were applied to determine the best fit solution. A comparison of modeling results is included in Appendix F.

The best fit solution for this aquifer is the Hantush No-Storage Method (1955), which models a leaky, semi-confined aquifer system. In this case, Zone III is the aquifer and the less permeable Zones I and II create a confining leakage response. AqtesolvTM solution printouts for the Hantush No-Storage Method, and the other methods used, are included in Appendix F. Calculations for leakage, aquitard, and aquifer characteristics are also included in Appendix F. The following parameters were established:

- $T = 122 \text{ ft}^2/\text{day};$
- $S = 1.49 \times 10^{-3}$; and
- K = 6.1 ft/day.

These parameters were used to calculate capture zone boundaries for withdrawal wells pumping at yields of 2 gpm. The capture zone calculations are included in Appendix F. The capture zones of the proposed groundwater withdrawal wells are shown on Figure 13. The withdrawal well system is described in detail in Section 10.6.

10.5.7.2 Assessment of Soil Corrective Measures Options

Only one soil sample in the former parts cleaning vat area, LS-1 at 410 mg/kg, had excessive TCE concentrations. Sample LS-1 was collected above the concrete floor of the former vat. The vertical Geoprobe profiles confirmed that no DNAPLs are present. There is no evidence of excessively impacted soil outside the former parts cleaning vat concrete vault.

Due to the extremely limited extent of soil contamination underneath the building, the shallow groundwater depth, and the unlikelihood that the unsaturated soil is acting as a significant source of further groundwater contamination, the corrective action plan proposes to remediate this area using an aggressive groundwater withdrawal system immediately downgradient of the Locomotive Shop building. The groundwater flushing effects created by the aggressive groundwater withdrawal in this area will effectively remove the limited soil contaminant concentrations. This proposal is outlined in Section 10.6.

If the Locomotive Shop building is destroyed, CSXT will thoroughly sample and remediate, if necessary, the subsurface soils underlying the former building.

APPENDIX F

AQUIFER TEST DATA ANALYSIS
METHOD	WELL	T (ft ² /min)	S	COMMENTS
Hantush	MW-1D	0.0596	2.88 x 10 ⁻³	<u>Leaky</u> aquifer solution excellent fit after $T = 1.5$ min. Possible
	MW-2D	0.0795	1.02 x 10 ⁻³	influence from pumping well for very early data. Measurement accuracy early not as good as later. Variability possible due to highly
	MW-3D	0.1163	5.57 x 10 ⁻⁴	variable aquifer.
Papadopolus-	MW-1D	0.309	7.51 x 10 ⁻⁶	<u>Confined</u> aquifer solution. Good fit, but aquifer not expected to act
Cooper	MW-2D	0.447	1.22 x 10 ⁻⁵	confined. T is a little higher than other solution. S is totally unreasonable and inconsistent. Calls into question the validity of
·	MW-3D	0.616	1.12 x 10 ⁻⁵	method.
Neuman	MW-1D	0.181	4.0 x 10 ⁻⁴	<u>Unconfined, delayed yield</u> solution. Marginal fit; however, aquifer may have delayed yield type response.
Theis (early	MW-1D	0.0877	2.52 x 10 ⁻³	<u>Confined</u> aquifer solution. Reasonably good fit for early data. Graph
data)	MW-2D	0.116	1.17 x 10 ⁻³	shows departure of later data that is typical for leaky aquifer.
MW-3D	0.175	6.41 x 10 ⁻⁴		
Moench	MW-1D	0.162	1.36 x 10 ⁻⁴	Leaky aquifer solution. Excellent fit. Method not strictly applicable to
	· MW-2D	0.223	1.73 x 10 ⁻⁴	this situation, but aquifer may act this way due to complex, interbedded materials. More distant wells have values significantly
	MW-3D	0.3698	3.70 x 10 ⁻⁵	different than other methods.
Jacob Method (distance drawdown)	MW-1, 2, and 3D	0.131	9.1 x 10 ⁻³	<u>Confined</u> aquifer solution. Excellent fit. Values are reasonable.

Conclusion: Hantush Method provides best fit and fairly consistent results in light of the complex and highly variable geology. Theis early data also has reasonable fit and reasonable results. Jacob Method has excellent fit and reasonable results. Other methods give poor fits, unreasonable results, or are not applicable. Recommend using Hantush Method results to model aquifer response.

Summary:	Hantush	$T avg = 0.0851 ft^{2}/min$	$S avg = 1.49 \times 10^{-3}$
	Theis (Early)	$T avg = 0.126 ft^{2}/min$	$S avg = 1.44 \times 10^{-3}$
	Jacob	<u>T avg = 0.131 ft²/min</u>	$S avg = 9.1 \times 10^{-3}$
56058-83	All	T avg = $0.114 \text{ ft}^2/\text{min}$	$S avg = 4.01 \times 10^{-3}$



MV HANTUSH NO-STORAGE **METHOD (1955)** CSX WAYCROSS 1 TITH = 0.07948 ft²/min ៵៹ <u>៴៴៹៝៶៲^៰៵៰៶៸៰៓៶៸៰៓៹៶៰៵</u> S = 0.001019 r∕B≈ Ø.5127 (ft)0.1 Drawdown 0.01 AQTESOLV GERAGHTY & MILLER, INC. Modeling Group 0.001 11111 1. 10. 100. 1000. 10000. Time (min)



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JACOB METHOD CONFINED SOLUTION



EVALUATION OF LEAKAGE AND AQUITARD CHARACTERISTICS FROM LOCOMOTIVE SHOP AQUIFER TEST DATA

Drawdown data was analyzed using a Hantush No-Storage Method (1955). From Aqtesolv output, the following characteristics were determined:

Well	T (ft ² /min)	S	r/B	r (ft)	B (ft)
MW-1D	0.0596	2.88 x 10 ⁻³	0.358	. 13	36
MW-2D	0.0795	1.02 x 10 ⁻³	0.513	40	78
MW-3D	0.1163	5.57 x 10 ⁻⁴	0.522	78	149

 $\overline{T} = 0.0851 \ ft^2/min = 122 \ ft^2/day$ $\overline{S} = 1.49 \ x \ 10^{-3}$ $\overline{B} = 88 \ ft$

LeakageFactor: $B = \sqrt{KDc}$

KD = T in ft/min

 $c = \frac{D'}{K'}$ = hydraulic resistance in minutes

D' = estimated thickness of aquitard in feet

K' = hydraulic conductivity of aquitard for vertical flow in ft/min

 $\overline{B^2} = \overline{KD} \ \overline{c} = \overline{T} \ \overline{c}$

$$\overline{c} = \frac{B^2}{\overline{T}} = \frac{(88 \ ft)^2}{(0.0851 \ ft.^2/\text{min})} = 91,000 \ \text{min} = 63.2 \ day$$

$$\frac{1}{c} = \frac{K'}{D'} = 0.016 \ day^{-1}$$

Assume D = Aquifer Thickness = 20 ft.Assume H = H' = Head in Aquifer = 24 ft.Assume $K = \frac{T}{D} = \frac{122 \text{ ft}^2/\text{day}}{20 \text{ feet}} = 6.1 \text{ ft/day} = 45.6 \text{ gpd/ft}^2$

Gradient = 0.005 ft/ft

EVALUATION OF LEAKAGE AND AQUITARD CHARACTERISTICS FROM CSXT PUMPING TEST DATA (Continued)

 $\overline{Q} = 2.09 \ gpm = 0.279 \ ft^3/min = 402 \ ft^3/day$

K' = Aquitard Hydraulic Conductivity = 0.128 ft/day = 0.000089 ft/min

D' = Aquitard Thickness = 8 ft







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	ARCADIS G&M of North Carolina, Inc. 801 Corporate Center Drive, Suite 300 Raleigh, NC 27607 Tel: 919–854–1202 Fax: 919–854–5448 www.arcadis-us.com	NOTES:			CSX TRANSPORTATION, INC. WAYCROSS YARD WAYCROSS, GEORGIA	CROS LF

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	PRJT MANAGER: J. BECKNER	CHECKED BY:	DRAFTER: A. WARREN	PREJECT NUMBER: SC000239.0006.00002	
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NOTES TO USERS

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NCS Information Services NOAA, NNRG512 National Geodatic Survey SSMC-3, AR020 1316 Eact-Vest Highway Steve Spang, Maryland 20910-3282 (301) 713-3242

To obtain surrent elevation, description, and/or location information for bench marks shown on this map, please contact the information Services Branch of the Natural Geodetic Survey at (301) 713-3242 or visit its website at http://www.ngs.ngia.gov/

Base map information shown on this FIRM was derived from National Agriculture imagery Program (ISAIP) at a scale of 1:20.000 from protography dated 2007 or later.

The profile base lines depicted on the map represent the hydraulic modeling baselines that metch the food profiles in the FIS report. As a result of improved biologistic data, the profile base line, in some cases, may devide significantly from the channel contentine or appear putside the SFHA.

Corporate limits shown on this map are based on the best data evaluable a The time of publication. Because changes die to annexations or de-annexations may have socured after this may was published may uses should contact appropriate community officials to verify current corporate limit

Please refer to the separately printed Map ladex for an overview map of the county shywing the tapout of map panels, community map repository addresses, and a Lating of Communities table containing National Flood Heaurence Program dates for each community as well as a tasting of the panels on Writch each community is scend.

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V you have questions about this map or questions concerning the National Rood insurance Program in general, please call 1-877-FEMA MAP (1-677-136-2827) or vali the FEMA website at http://www.lema.gov/businesahifo/

