GROUNDWATER QUALITY IN GEORGIA
FOR 2015

John C. Donahue and Anthony W. Chumbley

GEORGIA DEPARTMENT OF NATURAL RESOURCES
ENVIRONMENTAL PROTECTION DIVISION
WATERSHED PROTECTION BRANCH
WATERSHED PLANNING AND MONITORING PROGRAM

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CHAPTER 1 INTRODUCTION

1.1 PURPOSE AND SCOPE

This report, covering the calendar year 2015, is the twenty-ninth of the Circular 12 series. The first 19 reports, Circulars 12A through 12S, summarized the chemical quality of groundwater statewide across Georgia and utilized a static array of sampling stations that were sampled periodically, typically on a semiannual, annual, or biennial basis. The next five reports, Circulars 12T through 12X, dealt with specialized chemical groundwater quality issues: water quality in the Coastal region, water quality available to small public water systems, water quality in the Piedmont/Blue Ridge physiographic province, groundwater uranium in Georgia, and groundwater arsenic in Georgia. With this report and its predecessors, Circular 12Y, 12Z, 12AA and 12AB, monitoring the chemical quality of groundwater continues using a static array of periodically sampled stations.

These summaries are among the tools used by the Georgia Environmental Protection Division (EPD) to assess trends in the quality of the State's groundwater resources. EPD is the State organization with regulatory responsibility for maintaining and where possible, improving groundwater quality and availability. EPD has implemented a comprehensive statewide groundwater management policy of anti-degradation (EPD, 1991; 1998). Four components comprise EPD's current groundwater quality assessment program:

1. The Georgia Groundwater Monitoring Network. EPD's Watershed Protection Branch, Source Water Assessment Program, took over the Georgia Groundwater Monitoring Network from the Regulatory Support Program when that program disbanded in 2012. The Monitoring Network is designed to evaluate the ambient groundwater quality of eight aquifer systems present in the State of Georgia. The data collected from sampling of the Groundwater Monitoring Network form the basis for this report.

2. Water Withdrawal Program (Watershed Protection Branch, Water Supply Section). This program provides data on the quality of groundwater that the residents of Georgia are using.

3. Groundwater sampling at environmental facilities such as municipal solid waste landfills, Resource Conservation Recovery Act (RCRA) facilities, and sludge disposal facilities. The primary agencies responsible for monitoring these facilities are EPD's Land Protection and Watershed Protection Branches.
4. The Wellhead Protection Program (WHP), which is designed to protect areas surrounding municipal drinking water wells from contaminants. The United States Environmental Protection Agency (EPA) approved Georgia's WHP Plan on September 30, 1992. The WHP Plan became a part of the 1 Georgia Safe Drinking Water Rules, effective July 1, 1993. The protection of public supply wells from contaminants is important not only for maintaining groundwater quality, but also for ensuring that public water supplies meet health standards.

Analyses of water samples collected for the Georgia Groundwater Monitoring Network during the period January 2015 through December 2015 and from previous years form the database for this summary. The Georgia Groundwater Monitoring Network is presently comprised of 129 stations, both wells and springs. Twenty-one of the stations are scheduled for quarterly sampling; the remainder are scheduled to be sampled yearly. Each sample receives laboratory analyses for chloride, sulfate, nitrate/nitrite, total phosphorus, 26 metals, and volatile organic compounds (VOCs). Samples from the mineral spring and main well at Indian Springs State Park (stations P12A and P23) also receive analysis for fluoride. Field measurements of pH, conductivity, and temperature are performed on the sample water from each station. Field dissolved oxygen measurements are made on sample water from wells.

During the January 2015 through December 2015 period, Groundwater Monitoring staff collected 192 samples from 121 wells and 8 springs. A review of the data from this period and comparison of these data with those for samples collected for preceding monitoring efforts indicated that groundwater quality at most of the 129 stations has remained good.

1.2 FACTORS AFFECTING CHEMICAL GROUNDWATER QUALITY

The chemical quality of groundwater is the result of complex physical, chemical, and biological processes. Among the more significant controls are the chemical quality of the water entering the groundwater flow system, the reactions of the infiltrating water with the soils and rocks that are encountered, and the effects of the well-and-pump system.

Most water enters the groundwater system in upland recharge areas and in areas of leakage from adjacent geologic units. Water seeps through interconnected pore spaces and fractures in the soils and rocks until discharged to a surface water body (e.g., stream, lake, or ocean). The initial water chemistry, the amount of recharge, and the attenuation capacity of soils have a strong influence on the quality of groundwater in recharge areas. Chemical interactions between the water and the aquifer host rocks have an increasing significance with longer residence times. As a result, groundwater from discharge areas tends to be more highly mineralized than groundwater in recharge areas.
The well-and-pump system can also have a strong influence on the quality of the well water. Well casings, through compositional breakdown, can contribute metals (e.g., iron from steel casings) and organic compounds (e.g., tetrahydrofurans from PVC pipe cement) to the water. Pumps can aerate the water being drawn up and discharged. An improperly constructed or failing well can offer a conduit that allows local pollutants to enter the groundwater flow system.

1.3 HYDROGEOLOGIC PROVINCES OF GEORGIA

This report defines three hydrogeologic provinces by their general geologic and hydrologic characteristics (Figure 1-1). These provinces consist of:

1. The Coastal Plain Province of south Georgia;

2. The Piedmont/Blue Ridge Province, which includes all but the northwestern corner of north Georgia;

3. The combined Valley and Ridge and Appalachian Plateau Provinces of northwest Georgia.

1.3.1 Coastal Plain Province

Georgia’s Coastal Plain Province generally comprises a wedge of loosely consolidated sediments that gently dip and thicken to the south and southeast. Groundwater in the Coastal Plain flows through interconnected pore space between grains and through solution-enlarged voids in rock.

The oldest outcropping sedimentary formations (Cretaceous) are exposed along the Fall Line (Figure 1-1), which is the northern limit of the Coastal Plain Province. Successively younger formations occur at the surface to the south and southeast.

The Coastal Plain of Georgia contains a number of confined and unconfined aquifers. Confined aquifers are those in which the readily permeable layer of aquifer medium is interposed between two layers of poorly permeable material (e.g. clay or shale). If the water pressure in such an aquifer exceeds atmospheric pressure, the aquifer is artesian. Water from precipitation and runoff enters the aquifers and aquifer systems in their updip outcrop areas, where permeable sediments hosting the aquifer are exposed. Water may also enter the aquifers downdip from the recharge areas through leakage from overlying or underlying aquifers. Most
Figure 1-1. The Hydrogeologic Provinces of Georgia
Coastal Plain aquifers are unconfined in their updip outcrop areas, but become confined in downdip areas to the south and southeast, where they are overlain by successively younger rock formations. Groundwater flow through confined Coastal Plain aquifers is generally to the south and southeast, in the direction of dip of the sedimentary layers.

The sediments forming the major aquifer systems in the Coastal Plain range in age from Cretaceous to Holocene. Horizontal and vertical changes in the sediment layers that form these aquifer systems determine the thickness and extent of the aquifer systems. Several aquifer systems may be present in a single geographic area forming a vertical "stack".

The Cretaceous and Jacksonian aquifer systems (primarily sands) are a common source of drinking water within a 35-mile wide band that lies adjacent to and south of the Fall Line. However, the aquifer systems do extend downdip of the band. A well has been planned to test the Cretaceous aquifer along the Atlantic Coast for water supply development. Southwestern Georgia relies on three vertically stacked aquifer systems plus the upper part of the Cretaceous aquifer system for drinking water supplies: the Clayton, the Claiborne, and the Floridan aquifer systems. The Miocene/Surficial aquifer system (primarily sands) is the principal shallow aquifer system occupying much of the same broad area occupied by the Floridan aquifer system in central and eastern Georgia. The system is unconfined over most of its inland extent, but becomes partly confined both in the coastal area and in Grady, Thomas, Brooks, and Lowndes County area of South Georgia.

1.3.2 Piedmont/Blue Ridge Province

Though the Piedmont and Blue Ridge Physiographic Provinces differ geologically and geomorphologically, the two physiographic provinces share common hydrogeological characteristics and thus can be treated as a single hydrogeologic province. A two-part aquifer system characterizes the Piedmont/Blue Ridge Province (Daniel and Harned, 1997). The upper part of the system is the regolith aquifer, composed of saprolite and overlying soils and alluvium. The regolith aquifer is unconfined, and the water resides primarily in intergranular pore spaces (primary porosity). The lower aquifer in the Piedmont/Blue Ridge aquifer system is the bedrock aquifer. This aquifer is developed in metamorphic and igneous bedrock (mostly Paleozoic and Precambrian in age); the water resides in fractures and, in the case of specialized rocks such as marbles, solution-enlarged voids (secondary porosity). In contrast to the regolith aquifer, no intergranular (primary) porosity exists in the bedrock aquifer. The bedrock aquifer is semi-confined, with the overlying regolith aquifer media offering local confinement. The regolith aquifer also serves as the reservoir that recharges the bedrock aquifer.

1.3.3 Valley and Ridge Province

Faulted and folded consolidated Paleozoic sedimentary formations characterize the Valley and Ridge Province. The principal porosity present in aquifer media consists of fractures and solution-enlarged voids in the carbonate rocks; intergranular porosity
may be important in some places. Locally, groundwater and surface-water systems closely interconnect. Dolostones and limestones of the Knox Group are the principal aquifers where they occur in fold axes at the centers of broad valleys. The greater hydraulic conductivities of the thick carbonate sections in this province permit higher yielding wells than in the Piedmont/Blue Ridge Province.

1.3.4 Appalachian Plateau Province

Rocks in this province consist of consolidated Paleozoic sediments inclusive of the Mississippian and Pennsylvanian. Faulting and folding are less intense than in the Valley and Ridge province, and sediments tend to be flatter lying and more continuous areally. As in the Valley and Ridge Province, secondary porosity is the most important type of porosity. The highly fractured Fort Payne Chert and the Knox Group are major water-bearing units in this province.

Only a small part of this province extends into Georgia, at the State’s far northwest corner (Dade County and parts of Chattooga and Walker Counties). Due to its small extent in Georgia and its lack of monitoring stations for the current project, the Appalachian Plateau Province is combined with the Valley and Ridge Province for the purposes of this report.

1.4 REGIONAL GROUNDWATER PROBLEMS

Data from groundwater investigations in Georgia, including those from the GroundWater Monitoring Network, indicate that virtually all of Georgia has shallow groundwater sufficient for domestic supply. Iron, aluminum, and manganese are the only constituents that occur routinely in concentrations exceeding drinking water standards. These metals are mostly naturally occurring and do not pose a health risk. Iron and manganese can cause reddish or yellowish-brown to dark brown or black stains on objects and can give water a bitter metallic taste. Aluminum can cause water to appear cloudy.

In the karstic carbonate terranes of the combined Valley and Ridge/Appalachian Plateau Province, interconnection between the surface water systems and the groundwater systems can be extensive enough such that waters supplying some wells and springs (e.g., Crawfish Spring and Cedartown Spring) have been deemed under direct surface influence, requiring surface water type treatment if used for public supplies.

In the Piedmont/Blue Ridge Province, water available to wells drilled into bedrock consisting of granitic intrusive rocks, granitic gneisses, or hornblende gneiss/amphibolite assemblages occasionally may contain excessive naturally occurring uranium.
Aquifers in the outcrop areas of Cretaceous sediments south of the Fall Line typically yield acidic water that may require treatment. The acidity occurs naturally and results from the inability of the sandy aquifer sediments to neutralize acidic rainwater and from biologically influenced reactions between infiltrating water and soils. Groundwater from the Cretaceous along the coast is typically brackish.

Nitrate/nitrite concentrations in shallow groundwater from the farm belt in southern Georgia are usually within drinking water standards, but are somewhat higher than levels found in other areas of the State.

Three areas of naturally reduced groundwater quality occur in the Floridan aquifer system. The first is the karstic Dougherty Plain of southwestern Georgia. The second is the Gulf Trough area. The third is in the coastal area of east Georgia.

In the Dougherty Plain, as with the carbonate terranes of northwestern Georgia, surface waters and the contaminants they entrain can directly access the aquifer through sink holes.

The Gulf Trough is a linear geologic feature extending from southwestern Decatur County through northern Effingham County and may represent a filled-in marine current channel way (Huddleston, 1993). Floridan groundwater in and near the trough may be high in total dissolved solids and may contain elevated levels of sulfate, barium, radionuclides, and arsenic (Kellam and Gorday, 1990; Donahue et al., 2013).

In the Coastal area of east Georgia, the influx of water with high dissolved solids contents can dramatically raise levels of sodium, calcium, magnesium, sulfate, and chloride. In the Brunswick part of the Coastal area, groundwater withdrawal from the upper Floridan results in the upwelling of groundwater with high dissolved solids content from the deeper parts of the aquifer system (Krause and Clarke, 2001). In the Savannah portion of the Coastal area, heavy pumping in and around Savannah has caused a cone of depression which has induced seawater to enter the Floridan aquifer system and to flow down-gradient toward Savannah. The seawater has not yet reached Savannah and may not reach Savannah for many years. The seawater enters the aquifer system via breaches in the Miocene confining unit along the bottoms of waterways and sand-filled paleochannels offshore of the Beaufort/Hilton Head area of South Carolina (Foyle et al., 2001; Krause and Clarke, 2001).
CHAPTER 2 GEORGIA GROUNDWATER MONITORING NETWORK

2.1 MONITORING STATIONS

For the period January 2015 through December 2015, attempts were made to place sampling stations in the Coastal Plain Province's six major aquifer systems, in the Piedmont/Blue Ridge Province, and in the Valley and Ridge/ Appalachian Plateau Province (Table 2-1). Stations are restricted to wells or springs tapping a single aquifer or aquifer system. Attempts were made to have some monitoring stations located in the following critical settings:

1. areas of recharge;

2. areas of possible pollution or contamination related to hydrogeologic settings (e.g., granitic intrusions, the Dougherty Plain, and the Gulf Trough);

3. areas of significant groundwater use.

Most of the monitoring stations are municipal, industrial, and domestic wells that have well construction data.

2.2 USES AND LIMITATIONS

Regular sampling of wells and springs of the Groundwater Monitoring Network permits analysis of groundwater quality with respect to location (spatial trends) and time of sample collection (temporal trends). Spatial trends are useful for assessing the effects of the geologic framework of the aquifer and regional land-use activities on groundwater quality. Temporal trends permit an assessment of the effects of rainfall and drought periods on groundwater quality and quantity. Both trends are useful for the detection of non-point source pollution. Non-point source pollution arises from broad-scale phenomena such as acid rain deposition and application of agricultural chemicals on crop lands.

It should be noted that the data of the Groundwater Monitoring Network represent water quality in only limited areas of Georgia. Monitoring water quality at the 129 sites located throughout Georgia provides an indication of groundwater quality at the locality sampled and at the horizon corresponding to the open interval in the well or to the head of the spring at each station in the Monitoring Network. Caution should be exercised in drawing unqualified conclusions and applying any results reported in this study to groundwaters that are not being monitored.
<table>
<thead>
<tr>
<th>Aquifer or Aquifer System</th>
<th>Number of Stations Visited (Samples Taken)</th>
<th>Primary Stratigraphic Equivalents</th>
<th>Age of Aquifer Host Rocks</th>
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<tbody>
<tr>
<td>Cretaceous</td>
<td>25 stations (25 samples)</td>
<td>Ripley Formation, Cusseta Sand, Blufftown Formation, Eutaw Formation, Tuscaloosa Formation, Providence Sand, Steel Creek Formation, Gaillard Formation, Pio Nono Formation</td>
<td>Late Cretaceous</td>
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<tr>
<td>Clayton</td>
<td>3 stations (3 sample)</td>
<td>Clayton Formation</td>
<td>Paleocene</td>
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<td>Claiborne</td>
<td>3 stations (3 samples)</td>
<td>Claiborne Group</td>
<td>Middle Eocene</td>
</tr>
<tr>
<td>Jacksonian</td>
<td>8 stations (8 samples)</td>
<td>Barnwell Group</td>
<td>Late Eocene</td>
</tr>
<tr>
<td>Floridan</td>
<td>33 stations (63 samples)</td>
<td>Ocala Group, Suwanee Limestone</td>
<td>Middle Eocene to Early Oligocene</td>
</tr>
<tr>
<td>Miocene/Surficial</td>
<td>7 stations (7 samples)</td>
<td>Hawthorne Group, Miccosukee Formation, Cypresshead Formation</td>
<td>Miocene to Recent</td>
</tr>
<tr>
<td>Piedmont/Blue Ridge</td>
<td>44 stations (74 samples)</td>
<td>Various igneous and metamorphic complexes</td>
<td>Precambrian and Paleozoic</td>
</tr>
<tr>
<td>Valley and Ridge/Appalachian Plateau</td>
<td>6 stations (9 samples)</td>
<td>Shady Dolomite, Knox Group, Conasauga Group</td>
<td>Paleozoic, mainly Cambrian, Ordovician</td>
</tr>
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Stations of the Groundwater Monitoring Network are intentionally located away from known point sources of pollution. The stations provide baseline data on ambient water quality in Georgia. EPD requires other forms of groundwater monitoring for activities that may result in point source pollution (e.g., landfills, hazardous waste facilities, and land application sites) through its environmental facilities permit programs.

Groundwater quality changes gradually and predictably in the areally extensive aquifer systems of the Coastal Plain Province. The Monitoring Network allows for some definition of the chemical processes occurring in large confined aquifers. Unconfined aquifers in northern Georgia and in the surface recharge areas of southern Georgia are of comparatively small extent and more open to interactions with land use activities. The wide spacing of most monitoring stations does not permit equal characterization of water-quality processes in these settings. The quality of water from monitoring stations drawing from unconfined aquifers represents only the general nature of groundwater in the vicinity of the stations. Groundwater in the recharge areas of the Coastal Plain aquifer systems is one of the future drinking-water resources for down-flow areas. Monitoring stations in these recharge areas, in effect, constitute an early warning system of potential future water quality problems in confined portions of the Coastal Plain aquifer systems.

2.3 ANALYSES AND DATA RETENTION

Analyses are available for 192 water samples collected from 129 stations (121 wells and 8 springs) during the period January 2015 through December 2015. In 1984, the first year of the Groundwater Monitoring Network, EPD staff sampled from 39 wells in the Piedmont/Blue Ridge and Coastal Plain Provinces. Between 1984 and 2004, the network had expanded to include 128 stations situated in all three hydrogeologic provinces, with most of the stations being in the Coastal Plain Province.

Groundwater from all monitoring stations is tested for chloride, sulfate, nitrate/nitrite, total phosphorus, a variety of metals, and VOCs. Water from stations P12A and P23 also receive testing for fluoride. Appendix Table A-9 lists the EPA methods used to test for these analytes along with a reporting limit for each analyte. The results of the chemical tests are reported in this Circular. Before collecting a sample, EPD personnel also observe certain field measurements – pH, conductivity, dissolved oxygen, and temperature. This Circular also reports these measurements.

Testing for aluminum, beryllium, calcium, cobalt, iron, potassium, magnesium, manganese, sodium, titanium, and vanadium was undertaken using the inductively-coupled plasma (ICP) method (EPA method 200.7 in Table A-9). This method works well for the mostly major metals listed above. This method was also used to test for arsenic, barium, cadmium, chromium, copper, nickel, lead, antimony, selenium, tellurium, and zinc. The inductively-coupled plasma mass spectrometry (ICPMS) method (EPA method 200.8 in Table A-9) was also used to test for the metals mentioned in the
previous sentence as well as for molybdenum, silver, tin, and uranium. The ICPMS method generally gives better results for trace metals.

Pursuant to the Georgia Safe Drinking Water Act of 1977, EPD has established Maximum Contaminant Levels (MCLs) for certain analytes and other parameters, certain of which are included in analyses performed on Groundwater Monitoring samples (EPD, 2009). Primary MCLs pertain to analytes that can adversely affect human health if the maximum concentration for an analyte is exceeded for drinking water. Secondary MCLs pertain to parameters that may give drinking water objectionable, though not health-threatening, properties that may cause persons served by a public water system to cease using the water. Unpleasant taste and the ability to cause stains are examples of such properties. MCLs apply only to treated water offered for public consumption; nevertheless, they constitute useful guidelines for evaluating the quality of untreated (raw) water. Table A-10 in the Appendix lists the Primary and Secondary MCLs for Groundwater Monitoring Network analytes.

Most wells currently on the Monitoring Network have in-place pumps. Using such pumps to purge wells and collect samples reduces the potential for cross-contamination that would attend the use of portable pumps. Two wells, the Miller Ball Park North East Well (PA9C) and the Springfield Egypt Road Test Well (MI17), are flowing, which dispenses altogether with pumps and lessens the effects of the pump-well system on sample water. The pump on the Murphy Garden Well (MI9A), a shallow bored well formerly used for garden watering, is now out of operation and a bailer is used for sampling.

Sampling procedures are adapted from techniques used by United States Geologic Survey (USGS) and EPA. For wells except PA9C, MI9A, and MI17, EPD personnel purge the wells (EPA recommends removing three to five times the volume of the water column in the well) before collecting a sample to reduce the influence of the well, pump, and plumbing system on water quality. A purge of 15 to 20 minutes is usually sufficient to allow readings of pH, conductivity, temperature, and dissolved oxygen to stabilize and to allow corrosion films on the plumbing to be flushed away.

The apparatus used for monitoring field measurements and collecting samples consists of a garden hose with two branches at its end and a container. One branch conveys water to a container; the other branch allows the water to flow freely. On the container branch, water enters the bottom of the container, flows past the probe of the instrument taking field measurements, and discharges over the top of the container. Such an apparatus minimizes the exposure of the sample water to atmosphere. Once the field measurements have stabilized, sample containers are then filled with water discharging from the end of the free-flowing branch. Sample waters do not pass through a filter before collection. As a rule, trends for field measurements with increasing purge time include a lowering of pH, conductivity and dissolved oxygen. For shallower wells, the temperature tends to approach the mean atmospheric temperature for the area. For deeper wells geothermal heating may become apparent.

2-4
Once the sample bottles are filled, they are promptly placed on ice to preserve water quality. EPD personnel transport samples to the laboratory on or before the Friday of the week during which the samples were collected, well before holding time for the samples lapse. Field measurements and analytical results are provided in Table A-1 in the Appendix.

Files at EPD contain records of the field measurements and chemical analyses. Owners of wells or springs receive copies of the laboratory analysis sheets as well as cover letters and laboratory sheet summaries. The cover letters state whether or not any MCLs were exceeded. The Drinking Water Program’s Compliance and Enforcement Unit receives notification of Primary MCL exceedances involving public water supplies.

Station numbering assigns each station a two-part alphanumeric designation, the first part consisting of an alphabetic abbreviation for the aquifer being sampled and the second part consisting of a serial numeral, sometimes with an alphabetic suffix, the two parts separated by a dash. Some wells were also added from previous sampling and monitoring programs that were previously labeled with a County alphabetic abbreviation instead of an aquifer. In this case the previous identification number was retained for cross reference with previous samples. In order for the groundwater database to be compatible with a Watershed Protection Branch-wide water database, the stations were also assigned a three-part alphanumeric designation, the first part being an alphabetic abbreviation “GW” (for groundwater), the second part representing the local river basin and the third part being numeric.
CHAPTER 3 CHEMICAL GROUNDWATER QUALITY IN GEORGIA

3.1 OVERVIEW

Georgia’s major aquifer systems are grouped into three hydrogeologic provinces for the purposes of this report: the Coastal Plain Province, the Piedmont/Blue Ridge Province, and the Valley and Ridge/Appalachian Plateau Province.

The Coastal Plain Province comprises six major aquifer systems that are restricted to specific regions and depths within the Province (Figure 3-1). These major aquifer systems commonly incorporate smaller aquifers that can be locally confined. Groundwater Monitoring Network wells in the Coastal Plain aquifer systems are generally located in three settings:

1. Recharge (or outcrop) areas that are located in regions that are geologically updip and generally north of confined portions of these aquifer systems;

2. Updip, confined areas that are located in regions that are proximal to the recharge areas, yet are confined by overlying geologic formations. These are generally south to southeast from the recharge areas;

3. Downdip, confined areas, located to the south or southeast in the deeper, confined portions of the aquifer systems, distal to the recharge areas.

The Piedmont/Blue Ridge Province comprises two regional aquifers, the regolith aquifer and the bedrock aquifer (Daniel and Harned, 1997). The regolith aquifer is composed of saprolite – bedrock that has undergone intense chemical weathering -- plus soil and alluvium. The regolith aquifer, highly porous and appreciably permeable, serves as the reservoir that recharges the bedrock. The igneous and metamorphic bedrock exhibits low porosity – nearly all of the porosity is secondary and consists of discontinuous fractures, but can be very permeable as fractures can locally transmit water rapidly. Despite the regional scale of these two aquifers, flow systems are small-scale and localized, in contrast to those of the Coastal Plain.

Paleozoic sedimentary formations characterize the combined Valley and Ridge/Appalachian Plateau Province, although unlike in the Coastal Plain, these sedimentary formations are consolidated and have been subjected to faulting and folding. Also in contrast to the Coastal Plain Province, the faulting and folding has resulted in the creation of numerous, small-scale localized flow systems in the Valley and Ridge/Appalachian Plateau Province. The major water-bearing units in the province are carbonate rocks. Faulting and fracturing of the carbonates has led to the widespread development of karst features, which significantly enhance porosity and permeability and exert a strong influence on local flow patterns.
Figure 3-1. The Major Aquifers and Aquifer Systems of the Coastal Plain Province (after Davis, 1990).
3.2 CRETACEOUS AQUIFER SYSTEM

3.2.1 Aquifer System Description

The Cretaceous aquifer system is a complexly interconnected group of aquifer subsystems developed in the late Cretaceous sands of the Coastal Plain Province. These sands crop out in an extensive recharge area immediately south of the Fall Line in west and central Georgia (Fig. 3-2). In east Georgia, overlying Tertiary sediments restrict Cretaceous outcrops to valley bottoms. Five distinct subsystems of the Cretaceous aquifer system, including the Providence aquifer, are recognized west of the Ocmulgee River (Pollard and Vorhis, 1980). These merge into three subsystems to the east (Clarke et al., 1985; Huddleston and Summerour, 1996). The aquifer thickens southward from the Fall line, where the clays and sands pinch out against crystalline Piedmont rocks, to a column approximately 2,000 feet thick at the southern limits of the main aquifer use area (limit of utilization, Figure 3-2).

The Providence aquifer, a prominent subsystem of the Cretaceous aquifer system in the western Coastal Plain, is developed in sands and coquina limestones at the top of the Cretaceous column. The permeable Providence Formation-Clayton Formation interval forms a single aquifer in the updip areas (Long, 1989) and to the east of the Flint River (Clarke et al., 1983). East of the Ocmulgee River, this joint permeable interval is termed the Dublin aquifer (Clarke et al., 1985). This report treats the Providence aquifer as a part of the Cretaceous aquifer system.

EPD used 25 wells to monitor the Cretaceous aquifer system. Reported depths ranged from 128 feet (K7) to 1025 feet (PD6). All except well MAC1, MAR1, K6 and K21 are local government owned public supply wells. Well MAR1 produces process water for a sand mining operation, well K6 produces process water for a kaolin mill and well K21 is a private residence well. All wells are sampled yearly.

3.2.2 Field Parameters

The pHs of sample waters from all 25 wells ranged from 3.99 (K9A) to 8.49 (TAL1), with a median of 5.29. As a rule, pHs of waters from the deeper wells are basic, while those from shallower wells are acidic. Well PD3 and TAL1 seem to be the exceptions. Their sampling pH of 8.33 (PD3) and 8.49 (TAL1) would be expected for a well about twice their reported depth of 456 feet (PD3) and 300 feet (TAL1). Conductivities are available for all 25 wells and ranged from 18 uS/cm (K19) to 368 uS/cm (PD3), with a median of 49 uS/cm. As a rule, the deeper wells gave water with the higher conductivities. The temperatures measured should be viewed as approximations of the temperature of the water in the aquifer. Temperatures over all 25 well samples ranged from 17.98 degrees C (WEB1) to 27.98 degrees C (K20). Comparing well depths with sample water temperatures shows that the deeper wells generally tend to yield water with higher temperatures. The water temperature can also depend somewhat on the time of year measured,
Figure 3-2. Locations of Stations Monitoring the Cretaceous Aquifer System.
since sample water must traverse a zone influenced by surface temperature on its way from the aquifer to the measurement point. Dissolved oxygen measurements are available for 23 of the 25 wells. Concentrations ranged from 1.15 mg/L (STW1) up to 11.56 mg/L (MAC1). Generally, the dissolved oxygen content of groundwater decreases with depth. Dissolved oxygen measurements can suffer from various interferences, processes that can expose the groundwater to air. An inadequately purged well may deliver water that has been in contact with air in the well bore. Pumping a well’s water level down near the pump intake can entrain air in the pumped water. Also, pumping the water level in the well below a recharging horizon allows water to “cascade” or fall freely down the well bore and splash, thereby becoming aerated.

3.2.3 Major Anions, Non-Metals, and Volatile Organic Compounds

Testing for chloride, sulfate, combined nitrate/nitrite, total phosphorus, and volatile organic compounds (VOCs) was done for samples from all 25 wells. None of the 25 samples contained detectable chloride. Sulfate was detected in samples from eight wells, with concentrations all at or below 42 mg/L. Nitrate/nitrite was detected in 15 samples and ranged up to 2.10 mg/L (GLA1). Samples from nine wells contained detectable phosphorus, with concentrations ranging up to 0.27 mg/L (MAC1). Chloroform was detected in well K7 at a concentration of 2.0 ug/L.

3.2.4 Metals by Inductively-Coupled Plasma Spectrometry (ICP)

All 25 samples contained detectable sodium, which ranged from 1,000 ug/L (K9A) to 82,000 ug/L (PD3). The current high reporting limit for analyzing potassium accounts for the lack of potassium detections. Five wells gave samples with detectable aluminum ranging up to 340 ug/L (K12). Sixteen wells yielded samples containing detectable calcium, and 16 wells gave samples containing detectable iron. Calcium levels ranged from undetected to 64,000 ug/L (WEB1). Iron levels ranged up to 2,200 ug/L (BUR2), with samples from ten wells exceeding the Secondary MCL of 300 ug/L. Seven samples contained detectable magnesium, with a maximum value of 4,000 ug/L (PD6). Ten wells gave samples with detectable manganese. None exceeded the Secondary MCL of 50 ug/L. Beryllium, cobalt, potassium, vanadium, and titanium remained undetected.

3.2.5 Metals by Inductively-Coupled Plasma Mass Spectrometry (ICPMS)

ICPMS analysis found detectable levels only of copper, zinc, lead, barium and chromium. Barium was detected in all 25 samples with a maximum concentration of 66 ug/L (CHT1). Copper was detected in samples from seven wells with the maximum level at 32 ug/L (K12); zinc was detected in samples from five wells, with the maximum level at 220 ug/L (BUR2); lead was detected in samples from eight wells, with the maximum level at 40.0 ug/L (K12). With the exception of well K12’s lead action level exceedance, the copper and lead levels fell below their respective action levels of 1,300 ug/L and 15 ug/L and zinc below its secondary MCL of 5,000 ug/L. The highest concentrations for these three metals
tend to occur in samples with the lowest pHs. These three metals commonly leach into sample water from plumbing and are not necessarily present naturally. Chromium was detected at a concentration of 5.0 ug/L from well BUR2.

### 3.3 CLAYTON AQUIFER

#### 3.3.1 Aquifer System Description

The Clayton aquifer system of southwestern Georgia is developed mainly in the middle limestone unit of the Paleocene Clayton Formation. Limestones and calcareous sands of the Clayton aquifer system crop out in a narrow belt extending from northeastern Clay County to southwestern Schley County (Figure 3-3). Aquifer thickness varies, ranging from about 50 feet in the outcrop area to 265 feet in southeastern Mitchell County (Clarke et al., 1984). Both the Flint River to the east and the Chattahoochee River, to the west are the areas of discharge for the aquifer in its updip extent. Leakage from the underlying Providence aquifer system and from overlying permeable units in the Wilcox Formation confining zone provides significant recharge in downdip areas (Clarke et al., 1984). As mentioned previously, permeable portions of the Clayton and Providence Formations merge to form a single aquifer in the updip area and east of the Ocmulgee River. East of that river these combined permeable zones are called the Dublin aquifer.

#### 3.3.2 Field Parameters

EPD sampled three wells annually to monitor the Clayton aquifer system. Wells SUM1 and SUM2 are public supply wells and well CT8 is a private well. These wells vary in depth from 80 feet (CT8) to 230 feet (SUM2). The sample waters had a pH range of 4.23 (SUM2) to 5.03 (SUM1), an electrical conductivity range of 40 uS/cm (CT8) to 220 uS/cm (SUM2), a temperature range of 18.93 degrees C (CT8) to 19.40 degrees C (SUM1) and a dissolved oxygen range of 5.43 mg/L (SUM2) to 9.42 mg/L (CT8).

#### 3.3.3 Major Anions, Non-Metals, and Volatile Organic Compounds

Testing for chloride, sulfate, combined nitrate/nitrite, total phosphorus, and volatile organic compounds (VOCs) was done for samples from all three wells. One sample contained detectable chloride at a concentration of 11 mg/L (SUM1). Sulfate was detected in one sample with a concentration of 70 mg/L (SUM2). Nitrate/nitrite was detected in all three samples and ranged up to 1.8 mg/L (SUM1). No Samples contained detectable phosphorus. Chloroform was detected in well SUM1 (disinfectant by-product possibly from a leaky check valve) at a concentration of 0.78 ug/L.
Figure 3.3: Location of the Stations Monitoring the Clayton Aquifer.
3.3.4 Metals by Inductively-Coupled Plasma Spectrometry (ICP)

All three samples contained detectable sodium, which ranged from 2,400 ug/L (SUM2) to 9,600 ug/L (SUM1). The current high reporting limit for analyzing potassium accounts for the lack of potassium detections. Two wells gave samples with detectable aluminum ranging up to 1,100 ug/L (SUM2). One well yielded a sample containing detectable calcium and two wells gave samples containing detectable iron. Calcium levels ranged from undetected to 16,000 ug/L (SUM2). Iron levels ranged up to 610 ug/L (SUM2), with this well being the only one to exceed the Secondary MCL of 300 ug/L. One sample contained detectable magnesium at a value of 7,500 ug/L (SUM2). All three wells gave samples with detectable manganese with one well (SUM2) exceeding the Secondary MCL of 50 ug/L. Beryllium, cobalt, potassium, vanadium, and titanium remained undetected.

3.3.5 Metals by Inductively-Coupled Plasma Mass Spectrometry (ICPMS)

ICPMS analysis found detectable levels only of copper, zinc, lead, barium and nickel. Barium was detected in all three samples with a maximum concentration of 96 ug/L (SUM2). Copper was detected in three samples with the maximum level at 15 ug/L (CT8); zinc was detected in samples from two wells, with the maximum level at 400 ug/L (SUM2); and lead was detected in samples from two wells, with the maximum level at 9.8 ug/L (SUM2). The copper and lead levels of all three wells fell below their respective action levels of 1,300 ug/L and 15 ug/L. Nickel was detected at a concentration of 11 ug/L in well SUM2.

3.4 CLAIBORNE AQUIFER

3.4.1 Aquifer Description

The Claiborne aquifer is developed primarily in the sandy units in the middle and lower portion of the Middle Eocene Claiborne Group of southwestern Georgia. Claiborne Group sands crop out in a belt extending from northern Early County through western Dooly County. Recharge to the aquifer occurs both as direct infiltration of precipitation in the recharge area and as leakage from the overlying Floridan aquifer system (Hicks et al., 1981; Gorday et al., 1997). The discharge boundaries for the updip portion of the aquifer are the Ocmulgee River to the east and the Chattahoochee River to the west. The aquifer generally thickens to the southeast and is more than 350 feet thick near its down dip limit of utilization (Figure 3-4) (Tuohy, 1984).

The clay-rich upper unit of the Claiborne Group, the Lisbon Formation, acts as a confining layer and separates the Claiborne aquifer from the overlying Floridan aquifer system (McFadden and Perriello, 1983; Long, 1989; Huddleston and Summerour, 1996). The lower, water-bearing parts of the group had been correlated with the Tallahatta Formation (e.g., McFadden and Perriello, 1983; Long, 1989: Clarke et al., 1996) or more recently, have been divided into two formations,
the upper one termed the Still Branch Sand and the lower one correlated to the Congaree Formation (Huddleston and Summerour, 1996). East of the Ocmulgee River, permeable Congaree-equivalent sands are included in the Gordon aquifer (Brooks et al., 1985).

Three stations, all in or near the recharge area, were available to monitor the Claiborne aquifer. Wells CL2 and CL4A are municipal public supply wells, and well CL8 supplies water for drinking and other purposes for a State forestry nursery. Well CL2 is 315 feet deep, CL4A is 230 feet deep, and CL8 is not known precisely, but is about 90 feet deep.

3.4.2 Field Parameters

The pHs of sample waters from two wells was mildly acidic (CL8 at 6.23 and CL4A at 6.51), while the third was mildly basic (CL2 at 7.44). Conductivities registered at 88 uS/cm (CL8), 168 uS/cm (CL4A), and 209 uS/cm (CL2); and temperatures registered at 19.66 degrees C (CL4A), 20.68 degrees C (CL2), and 20.75 degrees C (CL8). Dissolved oxygen contents measured at 1.03 mg/L (CL8) and 4.06 mg/L (CL2). Since well CL4A exposes water to air, there was no measurement for dissolved oxygen for the water at this well.

3.4.3 Major Anions, Non-Metals, and Volatile Organic Compounds

Well CL2 was the only station to give a sample with detectable nitrate/nitrite (0.45 mg/L as nitrogen). A sample from well CL4A contained detectable sulfate at 12 mg/L. Samples from two wells contained detectable phosphorus (CL4A at 0.34 mg/L and CL8 at 0.54 mg/L). None of the samples contained detectable chloride or VOCs.

3.4.4 Metals by Inductively-Coupled Plasma Spectrometry (ICP)

Calcium and sodium were detected in samples from all three wells. The maximum and minimum calcium concentrations were 44,000 ug/L (CL2) and 12,000 ug/L (CL8). The maximum and minimum sodium concentrations were 1,900 ug/L (CL8) and 1,700 ug/L (CL2). Detectable magnesium occurred only in the samples from well CL8 (1,300 ug/L) and CL4A (3,200 ug/L). Wells CL4A and CL8 gave samples with detectable iron at 2,100 ug/L and 570 ug/L respectively and manganese at 59 ug/L and 54 ug/L respectively. Both samples exceeded the iron and manganese Secondary MCLs of 300 ug/L and 50 ug/L respectively.

3.4.5 Metals by Inductively-Coupled Plasma Mass Spectrometry (ICPMS)

ICPMS analyses found barium in all three samples. The maximum and minimum barium concentrations were 38 ug/L (CL8) and 11 ug/L (CL2). The sample from well CL8 contained zinc at 15 ug/L, which was below any applicable MCLs or action levels. Well CL8 also registered the lowest pH.
Figure 3-4. Locations of Stations Monitoring the Claiborne Aquifer.
3.5 JACKSONIAN AQUIFER

3.5.1 Aquifer Description

The Jacksonian aquifer system (Vincent, 1982) of central and east-central Georgia is developed primarily in sands of the Eocene Barnwell Group, though isolated limestone bodies are locally important. Barnwell Group outcrops extend from Macon and Crawford Counties (Hetrick, 1990) eastward to Burke and Richmond Counties (Hetrick, 1992). Figure 3-5 shows the extent and most significant Jacksonian recharge areas. Aquifer sands form a northern clastic facies of the Barnwell Group; the sands grade southward into less permeable silts and clays of a transition facies (Vincent, 1982). The water-bearing sands are relatively thin, ranging from 10 to 50 feet in thickness. Limestones equivalent to the Barnwell Group form a southern carbonate facies and are included in the Floridan aquifer system. The Savannah River and the Flint River are the eastern and western discharge boundaries for the updip parts of the Jacksonian aquifer system. The Jacksonian aquifer system is equivalent to the Upper Three Runs aquifer, as discussed by Summerour et al. (1994), page 2, and Williams (2007), “General Hydrogeology” table.

Eight wells were available to monitor the Jacksonian aquifer system. Wells J1B and J8A are domestic wells, while all the other wells are public supply wells. All are drilled wells, and each is scheduled for yearly sampling.

3.5.2 Field Parameters

The pHs for all the wells were basic. The pHs range from 7.09 (J1B) to 7.65 (J4). Conductivities ranged from 278 uS/cm (WAS1) to 355 uS/cm (J5). Temperatures ranged from 16.40 degrees C for well J8A to 21.15 degrees C for well J5, with water from the deeper wells registering higher temperatures. Dissolved oxygen concentrations ranged from 1.21 mg/L for well J6 to 9.78 mg/L for well J4 and are usually lowest in the deeper wells.

3.5.3 Major Anions, Non-Metals, and Volatile Organic Compounds

Sample waters from wells J5 and J6 contained detectable sulfate of 12 mg/L and 13 mg/L respectively. Nitrate/nitrite was detected in five of the eight samples ranging from undetected to 2.5 mg/L as nitrogen (J1B), and all measurements were below the Primary MCL of 10 mg/L as nitrogen. Phosphorus was detected in water from seven of the eight wells and ranged from undetected to 0.15 mg/L (J6). No sample waters contained detectable chloride. Bromodichloromethane was detected in well J4 (disinfectant by-product possibly from a leaky check valve) at a concentration of 0.57 ug/L.
Figure 3-5. Locations of Stations Monitoring the Jacksonian Aquifer.
3.5.4 Metals by Inductively-Coupled Plasma Spectrometry (ICP)

All eight wells gave waters with detectable calcium from 50,000 ug/L (J4) to 64,000 ug/L (J5 and J8A). Magnesium was detected in seven of the eight wells and ranged from undetected in J1B to 2,600 ug/L (J5). Detectable sodium occurred in each well sample and ranged from 2,100 ug/L (J6) to 4,400 ug/L (J1B). Iron was detected in three of the eight wells and ranged from undetected to 180 ug/L (J6). Well J8A and JEF1 gave a sample containing 14 ug/L and 71 ug/L manganese respectively. The sample from well JEF1 exceeded the manganese Secondary MCL of 50 ug/L. According to Kellam and Gorday (1990), the high calcium/magnesium ratios for these wells signifies that they derive most of their recharge from local surface water.

3.5.5 Metals by Inductively-Coupled Plasma Mass Spectrometry (ICPMS)

Seven of the eight wells yielded waters containing detectable barium, with a range from undetected (JEF1) to 89 ug/L (WAS1). Analysis found no other trace metals.

3.6 FLORIDAN AQUIFER SYSTEM

3.6.1 Aquifer System Characteristics

The Floridan aquifer system is developed predominantly in Eocene and Oligocene limestones and dolostones that underlie most of the Coastal Plain Province (Figure 3-6). The aquifer is a major source of groundwater for much of its outcrop area and throughout its downdip extent to the south and east.

The upper water-bearing units of the Floridan are the Eocene Ocala Group and the Oligocene Suwanee Limestone (Crews and Huddlestun, 1984). These limestones and dolostones crop out in the Dougherty Plain (a karstic area in southwestern Georgia) and in adjacent areas along strike to the northeast. In parts of Camden and Wayne Counties, the Oligocene unit is absent and the upper portions of the Floridan are restricted to units of Eocene age (Clarke et al., 1990). The lower parts of the Floridan consist mainly of dolomitic limestone of middle and early Eocene age and pelletal, vuggy, dolomitic limestone of Paleocene age, but extend into the late Cretaceous in Glynn County. The lower portions of the Floridan are hydrologically connected with the upper parts but are deeply buried and not widely used except for some municipal and industrial wells in the Savannah area. From its updip limit, defined by clays of the Barnwell Group, the aquifer system thickens to well over 700 feet in coastal Georgia.
A dense limestone facies occupying the Gulf Trough locally limits groundwater quality and availability (Kellam and Gorday, 1990; Applied Coastal Research Laboratory, 2001). The Gulf Trough may be a filled marine-current channel extending across Georgia from southwestern Decatur County through northern Effingham County. The trough, active beginning in the early Eocene, had ceased operating and filled with sediment in the Miocene.

A groundwater divide separates a smaller southwestward flow regime in the Floridan aquifer system in the Dougherty Plain in southwestern Georgia from the larger southeastward flow regime characteristic for the aquifer system under the remaining part of Georgia's Coastal Plain. Rainfall infiltration in outcrop areas and downward leakage from extensive surficial residuum recharge the Dougherty Plain flow system (Hayes et al., 1983). The main body of the Floridan aquifer system, lying to the east, is recharged by leakage from Jacksonian aquifer and by rainfall infiltration in outcrop areas and in areas where overlying strata are thin. Significant recharge also occurs in the area of Brooks, Echols, Lowndes, Cook and Lanier counties where the Withlacoochee River and numerous sinkholes breach the upper confining units (Krause, 1979).

Monitoring water quality in the Floridan aquifer system made use of 33 wells, with 23 scheduled for sampling on a yearly basis and 10 on a quarterly basis. The total number of samples collected from the wells was 63. All 33 wells are drilled wells. Thirty wells are local-government-owned public supply wells. One well supplies industrial process water, one well is a former USGS test well, and the remaining well supplies water for a coastal marina. Depths range from 174 feet (PA25 municipal well) to 1,211 feet (PA9C test well).

3.6.2 Field Parameters

Measurements of pH are available for all samples from all 33 wells and ranged from 7.20 (PA27) to 8.24 (GLY4). The median pH is 7.79 and the mean is 7.77. Conductivities are also available for all the samples from all the wells and ranged from 163 uS/cm (PA41A) to 1880 uS/cm (PA9C) with a median of 311 uS/cm and a mean of 349 uS/cm. Temperatures are available for all sampling events and ranged from 15.86 degrees C for well PA17 to 25.88 degrees C for well THO2 with a median of 22.67 degrees C and a mean of 22.66 degrees C. The high temperatures reflect the geothermal effect of the deeper wells. Fifty five dissolved oxygen measurements are available from 28 wells. The available measurements range from 0.79 mg/L (PA9C) to 10.07 mg/L (PA17). No measurements were taken at well PA4, PA5 and PA14A because the raw water outlet will not permit the attachment of the usual sampling apparatus and exposes sample water to air.

3.6.3 Major Anions, Non-Metals, and Volatile Organic Compounds

Ten Floridan wells yielded 15 samples containing detectable chloride. Chloride concentrations ranged from undetected to 820 mg/L (PA9C). The measurement for well PA9C is more than 20 times the next highest concentration of
39 mg/L for well PA4. Well PA9C derives water from the lower part of the Floridan aquifer. Twenty nine samples from 17 wells gave samples containing detectable sulfate. Levels ranged from undetected to 280 mg/L (PA9C). Seventeen water samples from 8 wells contained detectable nitrate/nitrite. Concentrations ranged from undetected to 1.7 mg/L as nitrogen (PA25). There is a general tendency for shallower wells to give samples with higher levels of nitrate/nitrite. Nitrate/nitrite levels in the samples from each quarterly sampled well tend, as a rule, to be similar to one another. Phosphorus was detected in 35 samples from 24 wells. Phosphorus levels ranged up to 0.09 mg/L (PA20) as total phosphorus. Volatile organic compounds (VOCs), consisting entirely of trihalomethane compounds, were detected in six samples from three wells (PA17, PA23 and PA34A). These compounds typically arise as byproducts from disinfection and their presence can indicate the reflux of treated water back down a well or result from sterilizing well plumbing following maintenance. For well PA23, samples regularly register detectable trihalomethanes, suggestive of leaky valves allowing treated water back down the well. For the remaining wells, the occasional nature of trihalomethane detections suggests a maintenance related origin.

3.6.4 Metals by Inductively-Coupled Plasma Spectrometry (ICP)

ICP analyses found detectable levels of potassium, manganese, iron, calcium, magnesium, aluminum and sodium. Detectable potassium occurred in only one sample from one well (PA9C). Failure to find detectable potassium in other samples results from the insensitivity of the testing procedure, as indicated by the high reporting limit (5,000 ug/L) for the metal. Detectable manganese occurred in 17 samples from nine wells. The maximum concentration of 100 ug/L occurred in a sample from well PA34A. All four samples from quarterly-sampled well PA34A and a sample from annually sampled well PA18 exceeded the Secondary MCL of 50 ug/L. The manganese levels in the samples from each of the quarterly sampled wells vary within a restricted range. Wells giving samples with manganese detections seem clustered in two areas: one in the Cook-Irwin-Lanier County area and the other in the Candler-Emanuel-Jenkins-Telfair-Toombs County area. Iron was detected in 27 samples from 15 wells. Of these, three samples exceeded the Secondary MCL of 300 ug/L; annual wells PA9C (1,000 ug/L), GLY4 (1,200 ug/L) and GLY2 (800 ug/L). The iron contents of samples from three quarterly wells (PA29, PA34A and PA36) seemed to vary within restricted ranges. Detectable magnesium was found in all samples from all wells except for those from quarterly well PA25. Magnesium concentrations ranged up to 84,000 ug/L (well PA9C), with a mean of 15,098 ug/L and a median of 15,000 ug/L. Non-detections were not included. Well PA25 is a Floridan recharge area well. Kellam and Gorday (1990) have noted that Ca/Mg ratios are higher in groundwaters from Floridan recharge areas, as is the case with this well. Magnesium levels in samples from each quarterly well seem to vary within relatively narrow ranges. Calcium was detected in all samples from the 33 Floridan wells. Concentrations ranged from 20,000 ug/L (PA41A) to 110,000 ug/L (PA9C), with a mean of 39,190 ug/L and a median of 36,000 ug/L. For samples from quarterly wells, calcium concentrations seem to fall within a narrow range for each well. Ca/Mg ratios in Floridan well waters have
Figure 3-6. Locations of Stations Monitoring the Floridan Aquifer System.
already been mentioned. Aluminum was detected above the Secondary MCL of 50-200 ug/L in four samples from four wells, PA34A (64 ug/L), PA36 (83 ug/L), PA38 (130 ug/L) and GLY4 (78 ug/L). Sodium was also found in all sample waters from all 33 wells and ranged in concentration from 1,700 ug/L (PA27) to 420,000 ug/L (PA9C), with a mean of 17,498 ug/L and a median of 7,800 ug/L. Sodium concentrations generally increase with depth.

3.6.5 Metals by Inductively-Coupled Plasma Mass Spectrometry (ICPMS)

ICPMS analysis found the following detectable metals in the Floridan samples: copper, zinc, lead, arsenic, selenium, molybdenum and barium. Four samples from quarterly well PA23 registered arsenic detection below or equal to the Primary MCL (10 ug/L). The well has given intermittent samples with detectable arsenic before. Annual wells MI10B, PA14A and VR8 gave samples showing detectable selenium below the Primary MCL (50 ug/L). Four samples contained detectable copper, one from quarterly well PA36 and three from quarterly well PA14A. Unlike most other wells, quarterly well PA14A furnishes sample water through a small diameter copper tube. Annual well PA9C and GLY4 along with quarterly well PA14A gave samples with detectable zinc. Only quarterly well PA14A contained detectable lead. Copper and lead detections were below the action levels of 1,300 ug/L for copper and 15 ug/L for lead. The zinc concentration fell below the Secondary MCL of 5,000 ug/L. Twelve samples drawn from quarterly wells PA23, PA28 and PA56 contained detectable molybdenum. Well PA23 produced the sample with the highest concentration of 54 ug/L. All three wells are in the Gulf Trough area. Barium was detected in all samples from all wells and ranged in concentration from 4.1 ug/L (PA14A) to 220 ug/L (PA39), all below the Primary MCL of 2,000 ug/L. The mean concentration of barium was 85.4 ug/L and the median was 86 ug/L. Barium seems to be more abundant in samples from wells of 400 foot to 700 foot depth range.

3.7 MIOCENE/SURFICIAL AQUIFER SYSTEM

3.7.1 Aquifer System Characteristics

The Miocene/Surficial aquifer system is developed in sands of the Miocene Hawthorne Group and of the Pliocene Miccosukee and Cypresshead Formations of the Georgia Coastal Plain (Figure 3-7).

The Hawthorne Group covers most of the Coastal Plain and consists predominantly of sand and clay (Huddleston, 1988), although carbonate rocks and phosphorites may locally be significant (Huddleston, 1988; Clarke et al., 1990). Clarke et al., 1990, note that three sequences consisting of a basal dense phosphatic limestone layer, a middle clay layer, and an upper sand layer typify the Miocene section in the coastal area. The sand layers in the two lowermost of the sequences host the lower and upper Brunswick aquifers, which are included in the Miocene/Surficial aquifer system of this report.
The Cypresshead Formation overlies the Hawthorne Group in the Coastal area (from the Atlantic coast to about 45 miles inland) and consists, in updip areas, predominantly of fine to coarse-grained quartz sand and, in downdip areas, interbedded fine sand and clay (Huddleston, 1988). In the Coastal Plain of far south central and southwestern Georgia, the Miccosukee Formation overlies the Hawthorne Group (Huddleston, 1988).

The Miccosukee Formation consists predominantly of sand but contains some clay. The characteristic lithology consists of thin-bedded to laminated fine to medium sand with scattered layers or laminae of clay. Also included in the aquifer system are Pleistocene arkosic sands and gravels interbedded with clays and Holocene sands and gravels interbedded with muds. The upper part of the aquifer system is unconfined, whereas, the deeper parts of the system may be locally confined and under artesian conditions.

Seven annually sampled wells were used to monitor the Miocene/Surficial aquifer system. Wells M11, M12A, M19A and M10B are private domestic wells, well WAY1 is a public supply well for a mobile home park and M19A and M10B are no longer being used as drinking water sources. Well M16 is used for general purposes at a fire station. Well M17 originated as a geologic bore hole — a hole drilled for investigating bedrock — that became a flowing well. It is currently used both as a domestic water source and as an augmentation well for maintaining a pond. Wells M12A and M19A are bored wells. The remainder are drilled wells. Depths, actual or approximate, have been determined for all seven wells.

3.7.2 Field Parameters

The pHs of the sample waters from the seven wells used to monitor the Miocene/Surficial aquifer system ranged from 3.97 (well M12A) to 8.00 (well M16). Two of the seven wells sampled (M12A and M10B) produced acidic water. The remaining five wells gave basic water. The acidic water-yielding wells included the two shallowest, while the basic water-producing wells included the two deepest. Conductivities ranged from 123 uS/cm (M12A) to 317 uS/cm (M19A). Water temperatures ranged from 19.12 degrees C (M17) to 22.90 degrees C (M11). Dissolved oxygen data are available for five of the seven wells and range from 1.68 mg/L (WAY1) to 9.44 mg/L (M12A). Valid dissolved oxygen measurements cannot be made on well M19A and M117 since one must be sampled with a bailer and the other is exposed to air before sampling.

3.7.3 Major Anions, Non-Metals, and Volatile Organic Compounds

Chloride registered at 12 mg/L in samples from the two bored wells M12A and M19A. The sample from the deepest Miocene well (M16) provided the only sulfate detection at 34 mg/L. Nitrate/nitrite was detected in sample waters from the bored wells M12A and M19A, one of the two lying in the range of likely human influence (≥ 3.1 mg/L as nitrogen) (Madison and Brunett, 1984). The former well registered 7.3 mg/L as nitrogen and the latter 2.0 mg/L. Detectable phosphorus was found in
samples from five of the seven wells. The concentrations ranged from 0.02 mg/L (MI1) to 0.32 mg/L (MI10B). None of the samples contained detectable VOCs.

3.7.4 Metals by Inductively-Coupled Plasma Spectrometry (ICP)

Samples from all seven wells contained calcium, magnesium, and sodium. Calcium levels ranged from 4,800 ug/L (well MI2A) to 44,000 ug/L (well MI17). Magnesium levels ranged from 1,900 ug/L (well MI17) to 16,000 ug/L (well MI16). Sodium levels ranged from 4,400 ug/L (well MI9A) to 16,000 ug/L (well MI16). Potassium was detected in well MI2A (6,900 ug/L) and well MI9A (7,200 ug/L). Iron was detected in the sample from well WAY1 at 24 ug/L and well MI10B at 1,600 ug/L. This last value far exceeds the Secondary MCL for iron of 300 ug/L. Manganese was found in samples from four wells: MI9A (40 ug/L), MI10B (69 ug/L), MI17 (12 ug/L) and WAY1 (120 ug/L). The 69 ug/L and 120 ug/L levels exceed the Secondary MCL for manganese of 50 ug/L. The high iron and manganese levels in water from drilled well MI10B are the reason the residents ceased using the water for household purposes, i.e., cooking, drinking, and laundring. Aluminum was detected in well MI2A at a concentration of 190 ug/L above the Secondary MCL of 50-200 ug/L.

3.7.5 Metals by Inductively-Coupled Plasma/Mass Spectrometry (ICPMS)

ICPMS analyses found detectable copper, zinc, selenium, barium, and lead in the Miocene aquifer samples. All seven samples contained detectable barium, which ranged in concentration from 18 ug/L (MI1) to 150 ug/L (MI10B). The sample from drilled well MI10B contained selenium at a level of 14 ug/L. Selenium at detectable levels is rare in Georgia’s groundwater. Zinc was detected in samples from well MI1 (38 ug/L) and MI10B (20 ug/L). Detectable lead occurred in the sample from bored well MI2A at a concentration of 3.4 ug/L. The sample from bored well MI2A also contained copper at a level of 8.3 ug/L. The copper, lead, and zinc in the water samples were likely derived from plumbing. None of the metals exceeded applicable action levels (1,300 ug/L for copper and 15 ug/L for lead) or MCLs (5,000 ug/L Secondary for zinc).

3.8 PIEDMONT/BLUE RIDGE AQUIFER SYSTEM

3.8.1 Aquifer System Characteristics

The Piedmont/Blue Ridge aquifer system in Georgia is part of the Piedmont and mountain aquifer system that extends from New Jersey into Alabama (Daniel and Harned, 1997). The system is unconfined or semiconfined and is composed of two major hydrogeologic units: a) regolith and b) fractured igneous and metamorphic bedrock (Heath, 1980; Daniel and Harned, 1997). Figure 3-8 shows the extent of the system in Georgia.
Figure 3-7. Locations of Stations Monitoring the Miocene/Surficial Aquifer System.
The regolith hydrologic unit is comprised of a mantle of soil, alluvium in and near stream bottoms and underlying saprolite. Saprolite is bedrock that has undergone extensive chemical weathering in place. Downward percolating, typically acidic, groundwater leaches alkali, alkaline earth and certain other divalent metals from micas, feldspars, and other minerals composing the original rock, leaving behind a clay-rich residual material. Textures and structures of the original rock are usually well-preserved, with the saprolite appearing as a "rotten" version of the original rock. The regolith unit is characterized by high, mostly primary porosity (35% to 55%) (Daniel and Harned, 1998) and serves as the reservoir that feeds water into the underlying fractured bedrock. Though it can store a great deal of water, saprolite, owing to its clay content, does not give up its water very rapidly. Saprolite grades downward through a transition zone consisting of saprolite and partially weathered bedrock with some fresh bedrock into fresh bedrock.

The fractured bedrock hydrologic unit is developed in igneous and metamorphic rocks. In contrast to the regolith, the porosity in such rocks is almost totally secondary, consisting of fractures and solution-enlarged voids. In the North Carolina Piedmont, Daniel and Harned (1997) found 1% to 3% porosity typical for bedrock. Fractures consist of faults, breaks in the rock with differential displacement between the broken sections, and joints, breaks in the rock with little or no differential displacement Heath (1980). Fractures tend to be wider and more numerous closer to the top of the bedrock. Daniel and Harned (1997) noted that at a depth of about 600 feet, pressure from the overlying rock column becomes too great and holds fractures shut. Fracturing in schistose rocks consists mainly of a network of fine, hair-line cracks which yield water slowly. Fractures in more massive rocks (e.g. granitic rocks, diabases, gneisses, marbles, quartzites) are mostly open and are subject to conduit flow. Thus wells intersecting massive-rock fractures are able to yield far larger amounts of water than wells in schistose rocks or even wells in regolith. Fractures can be concentrated along fault zones, shear zones, late-generation fold axes, foliation planes, lithologic contacts, compositional layers, or intrusion boundaries.

Seventy four samples from 40 wells and four springs were used to monitor water quality in the Piedmont/Blue Ridge aquifer system. Thirty nine of these wells are drilled. Thirty three of the 40 wells are public supply wells, and the remaining seven are domestic. One of the 40 wells is bored (P33) and is in domestic use. Of the four springs, three (P12A, HAS2 and TOW1) are mineral springs at State parks, and the other spring (BR5) is a public supply source. The State park mineral spring P12A and the following wells are scheduled for sampling on a quarterly basis: P21, P23, P25, P32, P33, P34, P35, P37 and BR1B. Well P25 was added to the network on a quarterly basis, and per agreement with the State Park manager an annual filtered sample is to be collected in addition to the quarterly unfiltered ones. The remaining stations are sampled on a yearly basis. Where their depths are known, wells deriving water from the bedrock aquifer range in depth from 150 feet to 705 feet. Domestic bored well P33, the only well drawing from the regolith aquifer, is 47 feet deep.
Figure 3-8. Locations of Stations Monitoring the Piedmont/Blue Ridge Aquifer System.
3.8.2 Field Parameters

Seventy four pH measurements from all 44 stations are available for the Piedmont/Blue Ridge aquifer system. The pHs ranged from 4.22 (P38) to 8.02 (BAN1A). Twenty total samples were basic; all four samples from quarterly well P32, four samples from quarterly spring P12A, three samples from quarterly well P35, and one sample from annual wells P20, P24, BAN1A, BAR1, COU2, ELB2, MAD1, UPS1 and WAS3. The remaining samples were acidic, including all samples from quarterly regolith well P33. The mean pH was 6.61 and the median 6.63. Conductivity measurements are available for all 74 samples. Conductivities range from 15 uS/cm (spring TOW1) to 964 uS/cm (well P32). The mean conductivity was 227 uS/cm and the median was 176 uS/cm. Samples with the higher pHs generally tended to have higher conductivities and vice versa. Temperatures were available for all sampled waters and range from 11.16 degrees C (spring TOW1) to 21.62 degrees C (well COU3). The mean temperature was 17.74 degrees C and the median was 17.85 degrees C. Geothermally elevated temperatures are not readily apparent for the Piedmont/Blue Ridge. Latitude, ground elevation, and season appear to have more influence on the sampling temperature. Dissolved oxygen measurements are available for 64 of the 74 samples from 37 of 44 stations. The samples from quarterly spring P12A and annual springs BR5, HAS2, TOW1 and wells P39, BAR1 and COU2 received no dissolved oxygen measurements since exposure of the sample water to air can render the measurement inaccurate. Dissolved oxygen levels ranged from 0.73 mg/L for well COU3 to 11.53 mg/L for quarterly well P34. The 11.53 mg/L reading lies above the oxygen saturation level for the temperature at sampling (18.28 degrees C). This reading suggests free-falling (cascading) water in the well or entrainment of air at the pump intake due to a low pumping water level and does not reflect the actual oxygen level in the groundwater.

3.8.3 Major Anions, Non-Metals, and Volatile Organic Compounds

All samples received testing for chloride, sulfate, nitrate/nitrite, total phosphorus, and VOCs. Four samples each from spring P12A and well P23, both located at Indian Springs State Park, received testing for fluoride. Seven stations yielded 13 samples with detectable chloride: quarterly well P37 with all four samples; quarterly spring P12A with all four samples; and annual wells P30, P41, COU4, ELB2 and WAS3 with one sample each. Well P37 gave the sample with the highest level at 74 mg/L. Detectable fluoride occurred in all four samples from well P23 at levels of 1.1 mg/L to 1.2 mg/L. Detectable fluoride also occurred in all four samples from quarterly spring P12A at levels ranging from 4.4 mg/L to 4.7 mg/L. This last range of levels exceeds the Primary MCL of 4 mg/L for fluoride; the spring water from this station has consistently done so in the past. Historical fluoride levels have ranged from slightly above 4 mg/L to slightly above 5 mg/L. Sulfate was detected in 34 samples from six quarterly and eleven annual stations, with the highest concentration (340 mg/L) occurring in a sample from quarterly well P32. Spring P12A and quarterly wells P32, P37, P34, P21 and BR1B each have
sulfate values that vary within narrow ranges. Nitrate/nitrite was detected in 53 of 74 samples from 32 stations with a high concentration of 3.1 mg/L as nitrogen for annual well P30 lying in the range of likely human influence (≥ 3.1 mg/L as nitrogen) (Madison and Brunett, 1984). This level is well below the Primary MCL of 10 mg/L as nitrogen. Detectable phosphorus occurred in 56 samples from 35 stations, with the highest concentration of 0.19 mg/L being found for quarterly well P34. Phosphorus concentrations vary within narrow ranges within the quartets of samples from quarterly spring P12A and from quarterly wells P21, P23, P25, P33, P34 and P35. Detectable VOCs occurred in samples from wells P41 (chloroform 0.66 ug/L), COU4 (chloroform 0.64 ug/L and methyl tert-butyl ether 1.1 ug/L) and UPS1 (chloroform 1.4 ug/L).

3.8.4 Metals by Inductively-Coupled Plasma Spectrometry (ICP)

ICP analysis found detectable aluminum, calcium, iron, potassium, magnesium, manganese, and sodium. No beryllium, cobalt, titanium, or vanadium was detected. Calcium was found in all samples except springs HAS2 and TOW1. These two springs are located in FD Roosevelt State Park (HAS2) and Brasstown Bald Recreation Area (TOW1). The reason for no detectable calcium in these two springs is probably because these two springs flow through a homogeneous quartzite rock. The highest calcium levels (220,000 ug/L, 140,000 ug/L, 130,000 ug/L and 110,000 ug/L) occurred in the quarterly samples from well P32. The mean calcium concentration was 27.135 ug/L and the median concentration was 16,500 ug/L. As a rule, calcium levels of samples from each quarterly station tend to cluster closely. Magnesium was detected in 65 samples from 38 stations. Magnesium contents of sample waters ranged from not detected up to 34,000 ug/L (well P30). As with calcium, magnesium levels in samples from each quarterly well generally tend to cluster. All samples from the quarterly regolith well P33 and samples from annual bedrock wells P38 and BAN1A and annual springs BR5, HAS2 and TOW1 contained no detectable magnesium. Sodium was present in all samples and ranged from 1,000 ug/L in the sample from spring HAS2 to 43,000 ug/L in a sample from well P37. Sodium levels for each quarterly well have a general tendency to cluster. The mean sodium concentration was 12,997 ug/L and the median was 11,000 ug/L. Detectable potassium was found in all four samples from one station (well P35). The low sensitivity of the current laboratory testing procedure for potassium probably accounts for the apparent scarcity of this metal. Aluminum was detected in five samples from wells P25, P33 and P42. Well P42 registered the highest level at 160 ug/L. Aluminum levels exceeded the low limit of 50 ug/L for the Secondary MCL range but not the upper limit. Iron was detected in 35 samples from 20 wells and one spring, with a range from not detected up to 7,400 ug/L (well FRA1). This concentration exceeds the Secondary MCL for iron of 300 ug/L. Six other wells produced samples with an iron level equal to or greater than the Secondary MCL; P41 (3,200 ug/L), P37 (3,000 ug/L, 1,300 ug/L, 1,200 ug/L and 330 ug/L), COU3 (1,400 ug/L), COU1 (1,000 ug/L), P25 (670 ug/L), and MAD1 (660 ug/L). Manganese was detected in 43 samples from 21 wells and one spring, with a maximum concentration of 300 ug/L (well COU4). Twenty samples from wells P20, P21, P25, P35, P37, COU1, COU3, COU4, FRA1, HAS1, MAD1 and WAS3 equaled or exceeded the Secondary MCL of 50 ug/L.
3.8.5 Metals by Inductively-Coupled Plasma Mass Spectrometry (ICPMS)

ICPMS analysis of water samples detected the following metals: copper, zinc, molybdenum, tin, barium, lead and uranium. None of the following metals were found in detectable amounts: chromium, nickel, arsenic, selenium, silver, cadmium, antimony, and thallium. Molybdenum was detected in only one sample from well BAN1A. Tin was detected in only one sample from well MAD1. Copper occurred in 14 samples from 8 wells, with a maximum level of 30 ug/L in the sample from well P22. This sample also had one of the lowest pHs. All copper detections occurred in acidic waters, with the highest pH for a sample containing detectable copper registering at 6.93. No detectable copper occurred in neutral or basic waters. Zinc was detected in 23 samples from 16 wells, with the maximum level at 230 ug/L from well STE1. All zinc detections except for wells P20 (pH 7.69), P24 (pH 7.11) and BAR1 (pH 7.15) occurred in acidic waters. Lead was detected in six samples from six wells. All lead detections occurred in acidic water. All lead detections occurred with zinc or copper detections. Again, these three metals commonly leach into sample water from plumbing and are not necessarily present naturally. Barium, as elsewhere in the State’s groundwater, was a nearly ubiquitous trace metal, being detected in 70 samples from 40 wells and four springs. Three samples from quarterly spring P12A and one sample from quarterly well P32 contained no detectable barium. The maximum concentration was 220 ug/L from a sample from annual well P20. No samples exceeded the Primary MCL of 2,000 ug/L. Uranium was detected in six samples from five wells. Uranium detections were down from previous years due to the reporting limit of the lab going from the previous 1.0 ug/L to 10 ug/L. Uranium concentrations ranged from not detected up to 19.6 ug/L found in a sample from well P34. Granitic bedrock is present where these wells are drilled and is the most common bedrock type to host uraniferous water.

3.9 VALLEY AND RIDGE/APPALACHIAN PLATEAU AQUIFER SYSTEM

3.9.1 Aquifer System Characteristics

Since Georgia’s portion of the Appalachian Plateau Province extends over such a small area of the State, i.e., its northwestern corner, this report includes that province with the Valley and Ridge Province for purposes of discussion. Bedrock in the combined province is sedimentary, comprising limestones, dolostones, shales, siltstones, mudstones, conglomerates and sandstones (Figure 3-9).

Primary porosity in the province’s bedrock is low, leaving fractures and solution-enlarged voids as the main water-bearing structures. The bedrock in the province is extensively faulted and folded, conditions that have served to proliferate fracturing and to segment water-bearing strata into numerous local flow systems, in contrast to the expansive regional flow regimes characteristic of the Coastal Plain sediments. Fractures in limestones and dolostones can become much enlarged by solution, greatly increasing their ability to store water.
Zones of intense fracturing commonly occur in carbonate bedrock along such structures as fold axes and fault planes and are especially prone to weathering. Such zones of intense fracturing give rise to broad valleys with gently sloping sides and bottoms covered with thick regolith. The carbonate bedrock beneath such valleys presents a voluminous source of typically hard groundwater.

As in the Piedmont/Blue Ridge Province, the regolithic mantle of soil and residuum derived from weathered bedrock blankets much of the Valley and Ridge/Appalachian Plateau Province. The water table lying within the regolithic mantle yields soft water ("freestone" water) sufficient for domestic and light agricultural use (Cressler et al., 1976; 1979). The regolithic mantle also acts as a reservoir, furnishing water to the underlying bedrock, which supplies most of the useful groundwater in the province.

Monitoring water quality in the Valley and Ridge/Appalachian Plateau aquifers made use of four springs and two drilled wells (Figure 3-9). Springs VR2A, VR8 and VR10 are public supply springs. Spring VR3 is a former public supply spring now serving ornamental purposes in a public park. Well VR1 is a public supply well, and well VR6A is an industrial process water source. Spring VR8 is scheduled for quarterly sampling, while all the other stations are sampled on an annual basis. All stations tap carbonate bedrock aquifers.

3.9.2 Field Parameters

Sample water pHs ranged from 7.06 for spring VR10 to 7.61 for spring VR8. Conductivities ranged from 241 uS/cm (well VR1) to 302 uS/cm (spring VR8). Dissolved oxygen measurements are available for well VR1 (6.15 mg/L) and well VR6A (7.95 mg/L). Dissolved oxygen measurements were made on spring waters at or downstream of spring heads; however, due to atmospheric exposure at the spring heads, these measurements may not validly represent oxygen levels in the water prior to discharge. The temperature of sample waters from well VR1 was 16.36 degrees C and from well VR6A was 19.91 degrees C. For spring waters, contact with the surface environment may have altered actual water temperatures present at the spring heads, since water temperatures were measured downstream from the springheads.

3.9.3 Major Anions, Non-Metals, and Volatile Organic Compounds

Neither chloride, sulfate nor phosphorus was detected in any of the sample waters. Detectable nitrate/nitrite was present in all of the sample waters and ranged from 0.69 mg/L as nitrogen in spring VR8 to 1.80 mg/L as nitrogen in spring VR10. The sample from well VR6A was the only one to contain detectable VOCs. The compounds consisted of: 1,1-dichloroethylene at 2.1 ug/L (Primary MCL = 7 ug/L) and tetrachloroethylene at 2.3 ug/L (Primary MCL = 5 ug/L). These compounds, particularly the chlorinated ethylenes, are used primarily as solvents. The owner/user of well VR6A manufactures barium and strontium compounds and anthraquinone.
3.9.4 Metals by Inductively-Coupled Plasma Spectrometry (ICP)

ICP analysis found calcium, magnesium, and sodium in all samples and iron in four samples. Detectable iron was present in three of the four samples from spring VR8 (21 ug/L, 27 ug/L and 48 ug/L) and in the sample from spring VR10 at 62 ug/L, all at levels below the Secondary MCL of 300 ug/L. Neither manganese nor aluminum was detected in any of the samples. Calcium levels ranged from 29,000 ug/L from well VR1 and VR6A to 42,000 ug/L from spring VR2A. Magnesium levels ranged from 15,000 ug/L from springs VR2A and VR10 to 18,000 ug/L from wells VR1 and VR6A. Sodium levels ranged from 1,300 ug/L from spring VR3 to 4,400 ug/L from well VR6A.

3.9.5 Metals by Inductively-Coupled Plasma Mass Spectrometry (ICPMS)

ICPMS analysis found barium, selenium, lead and zinc. Detectable barium was present in all samples and ranged from 9.1 ug/L from well VR1 to 520 ug/L from well VR6A. All samples save the one from VR6A have barium levels below 100 ug/L. Well VR6A furnishes process water to a firm that manufactures barium and strontium compounds and is situated in an area that sees the mining and processing of barite. Selenium was detected at a level of 15 ug/L in one of the samples from quarterly spring VR8. Lead at a level of 2.5 ug/L and zinc at a level of 14.0 ug/L were found in the sample from spring VR10. A spigot in the treatment house near the spring head or related plumbing may have contributed these three metals. This spigot is the only source of untreated water from the spring.
Figure 3-9. Locations of Stations Monitoring the Valley-and-Ridge/Appalachian Plateau Aquifer System.
CHAPTER 4 SUMMARY AND CONCLUSIONS

EPD personnel collected 192 water samples from 121 wells and eight springs on the Groundwater Monitoring Network during the calendar year 2015. The samples were analyzed for VOCs, chloride, sulfate, nitrate/nitrite, total phosphorus, 15 trace metals by ICPMS analysis, and 11 major metals by ICP analysis. Waters from two neighboring stations in the central Piedmont received analyses for fluoride because one of the stations was known to produce water with excessive levels of fluoride. These wells and springs monitor the water quality of eight major aquifers and aquifer systems as considered for this report in Georgia:

- Cretaceous/Providence aquifer system,
- Clayton aquifer,
- Claiborne aquifer,
- Jacksonian aquifer
- Floridan aquifer system,
- Miocene/Recent aquifer system,
- Piedmont/Blue Ridge aquifer system,
- Valley and Ridge/Appalachian Plateau aquifer system.

4.1 PHYSICAL PARAMETERS AND pH

4.1.1 pH

The Cretaceous/Providence aquifer system, developed in Coastal Plain sands, furnished waters with the overall lowest pHs. This aquifer system featured only six of 25 wells yielding waters with basic pHs.

Not many stations were available to sample wells tapping the Clayton, Claiborne, or Jacksonian aquifers. However, the results are these: 1) Clayton – acidic – as expected for updip portions of the aquifer, downdip portions should be basic; 2) Claiborne – one basic, two acidic – one acidic-yielding well is shallow and updip in sands; basic-yielding well is deeper and probably penetrates some limey sand or limestone; 3) Jacksonian – all eight wells were basic – basic and neutral waters should be expected from limey sands.

The Floridan aquifer system, as might be expected for carbonate-rock aquifers, gave waters with mildly basic pHs. Waters from the Floridan are the most basic in pH of any in the study.
The Miocene aquifer system is developed in sands. However, these may include shelly detritus in some places (evident at surface excavations near well MI17 and at coastal well MI16). Dissolution of such detritus is capable of raising the pHs of groundwaters in such areas, giving water from this well a mildly basic pH. In places where such shelly matter is not available, waters emerge with low pHs, as at well MI2A.

Sample-water pHs in the Piedmont/Blue Ridge are generally mildly acidic, with 21 out of 74 sample measurements exceeding or equaling a pH of 7.00.

The Valley-and-Ridge/Appalachian Plateau sampling stations are all located in the Valley-and-Ridge sector. With carbonate rocks being the major aquifer media, samples from the sector would be expected to be mildly basic, which all nine samples taken in the sector were found to be basic with some samples close to neutral. In the past, some of these samples were found to be slightly acidic instead of all samples being basic. The seeming incidence of past acidic waters was probably due to a larger amount of typically acidic precipitation entering the springs’ flow systems than the carbonate bedrock can neutralize.

The very acidic pHs of the sample waters in the updip portions of the Jacksonian, Clayton, Claiborne, and, particularly, the Cretaceous/Providence can face metal plumbing with leaching and corrosion problems. Such waters may contain elevated or excessive, but not naturally occurring, levels of lead, copper, and zinc.

4.1.2 Conductivity

Conductivity in groundwaters from the sandy Cretaceous/Providence aquifer system seems to be highest for the deeper wells near the Chattahoochee River. Overall, conductivities are relatively low, in the range of lower tens of microsiemens.

Similar conductivities can be found in waters from the updip portions of the Clayton and Claiborne aquifers, where the media consist mostly of sand. For the Piedmont/Blue Ridge aquifer system, low conductivities could be associated with groundwaters hosted by quartzites or quartz veins. High conductivities may arise in waters in deep flow regimes where waters are long in contact with granitic and other reactive host rocks.

Conductivities of groundwaters in the Floridan and other carbonate rock aquifers are generally higher than those in siliceous rocks. This condition results from the dissolution of carbonate minerals, in cases augmented by dissolution of intergranular sulfate, where dissolved sulfate will also be present in the water.
4.1.3 Temperature

Groundwater temperatures measured under the current sampling procedure are only approximations of the actual groundwater temperature, as some heating can result from the action of pumping and heating or cooling can result from exposure to ambient surface conditions. Nevertheless, groundwaters from shallower wells in the northern part of the State are overall somewhat cooler than those from the southern part; and those from wells much deeper than about 400 to 500 feet show effects from geothermal warming.

4.2 ANIONS, NON-METALS AND VOCs

4.2.1 Chloride and Fluoride

Water samples receive testing for fluoride only at Piedmont/Blue Ridge stations P12A, a mineral spring and well P23, a nearby well. All four samples from spring P12A exceeded the Primary MCL for fluoride. Testing more stations for fluoride could provide a better base level assessment of fluoride contents in the State’s ambient groundwaters.

Chloride at currently detectable levels is not too common in ambient groundwaters. Abundance seems to be largest in the deeper Floridan waters, which had detections at ten out of 33 stations. The Floridan occurrences seem restricted to the Gulf Trough and Coastal areas, with the Coastal area sample from well PA9C giving the study’s only Secondary MCL exceedance for chloride. The Miocene/Surficial aquifer is the next most abundant with two of seven stations of less than 100 feet depth giving water with detectable chloride. Chloride is also relatively abundant in Piedmont/Blue Ridge waters, detected at seven out of 44 stations.

4.2.2 Sulfate

Sulfate is more widespread than chloride. Sulfate is more abundant in deeper waters, with the shallowest occurrence, aside from Piedmont/Blue Ridge mineral spring P12A, being 150 feet-deep well MAR1 in the Cretaceous aquifer. Sulfate seems more abundant in Floridan sample waters, detectable at 16 out of 33 stations. Sulfate is also abundant in the Piedmont/Blue Ridge, occurring in detectable amounts in waters from 17 of 44 stations. The Cretaceous aquifer yielded samples containing detectable sulfate in eight out of 25 stations. Jacksonian sample waters yielded two out of eight stations with detectable sulfate. The sample from Piedmont well P32 yielded the study’s highest overall sulfate content and a Secondary MCL exceedance. The lowest incidences of detectable sulfate were in the Miocene/Surficial at one of seven stations.
4.2.3 Nitrate/Nitrite

One hundred five (105) samples from 74 of the 129 stations sampled for this project contained detectable nitrate/nitrite. At least one sampling station drawing from each of the aquifers and aquifer systems discussed in this report gave a sample with detectable nitrate/nitrite. The combined substances are most widespread among the Valley and Ridge/Appalachian Plateau samples, where all stations gave samples containing detectable amounts. The combined substances are also widespread in Piedmont/Blue Ridge and Floridan waters. The four highest concentrations of nitrate/nitrite (7.3 mg/L at well MI2A, 3.7 mg/L at well WKE1, 3.6 mg/L at well P42 and 3.1 mg/L at well P30) occurred at Miocene/Surficial and Piedmont stations. All four samples exceeded the 3 mg/L (as nitrogen) level generally considered to indicate human influence (Madison and Brunett, 1984; Gaskin et al., 2003).

Since nitrate/nitrite, an oxidant, becomes depleted the farther water travels away from oxidizing, near-surface environments and into reducing ones, a rude inverse relation exists between the concentration of the combined substances and well depths. The nitrate/nitrite concentrations in Floridan samples illustrate this: the combined substances are undetected in wells deeper than about 650 feet and reach a maximum concentration of 1.7 mg/L in four of four samples from well PA25, 174 feet deep. The situation in the Piedmont/Blue Ridge is less straightforward, as mineral spring P12A lacks detectable nitrate/nitrite and well P24 at 700 feet gives water with a concentration of 0.26 mg/L.

4.2.4 Phosphorus

Analyses determine only total phosphorus; the method used (EPA Method 365.1) for testing cannot determine how the element is bound. There were three samples from three stations collected for the Claiborne, however this aquifer registered the highest mean phosphorus content of 0.29 mg/L. Of the more extensively sampled Piedmont/Blue Ridge and Floridan aquifer systems, the former registered a mean phosphorus content of 0.06 mg/L and the latter a content of 0.016 mg/L. The high phosphorus value for the Floridan was .09 mg/L and the high for the Piedmont/Blue Ridge was 0.78 mg/L, which was the highest value for all the aquifers. The apparent low phosphorus content occurred for the Valley and Ridge/Appalachian Plateau aquifer system with no detections.

4.2.5 Dissolved Oxygen

The measurement of dissolved oxygen contents is beset with some difficulties that can cause spurious values: instrument malfunction; aeration of well water due to cascading or to a pump's entraining air at low pumping water levels; measuring at spring pools or at sampling points that cannot be isolated from atmosphere. Nevertheless, measured dissolved oxygen generally decreases with well depth.
4.2.6 Volatile Organic Compounds

Volatile organic compounds (VOCs) were found in 13 samples from 10 wells (see Table 4-2). None exceeded their respective Primary MCLs. The trihalomethanes -- chloroform, bromodichloromethane, chlorodibromomethane, and bromoform -- were the most widely occurring of the VOCs. These compounds result from halogen-bearing disinfectants reacting with organic matter naturally present in the water. Two scenarios accompany the occurrence of the compounds. The first involves disinfection of the well and plumbing components incident to maintenance or repairs, as took place in 2012 with well PA44. The second scenario involves leaking check valves or foot valves that allow disinfectant-treated water to flow back down the well when pumps are off, as apparently happened with well PA23.

Well VR6A yielded water containing chlorinated ethylene compounds. Sample water from VR6A also contained detectable chlorinated benzene compounds. The former are used as solvents; in addition to solvent uses, the latter can be used as disinfectants, fumigants, pesticides, and starters for manufacturing other compounds. The owner of well VR6A, Chemical Products Corporation, manufactures barium and strontium compounds.

Well COU4 yielded water containing methyl tert-butyl ether (MTBE; 2-methoxy-2-methyl-propane), which has no MCL. An advisory range of 20 μg/L to 40 μg/L has preliminarily been set due to offensive taste and smell. The compound has been added to motor fuels as an oxygenate (promotes cleaner burning). That use is being curtailed due to the greater water solubility of the compound compared to other fuel components thus its heightened ability to contaminate groundwater. Data on the long-term health effects of the compound are sparse.

4.3 ICP METALS

Analysis using inductively coupled plasma spectrometry (ICP) works well for metals that occur in larger concentrations in groundwater samples. Samples in this study were not filtered, so the method measured analytes that occurred in fine suspended matter as well as those occurring as solutes. The laboratory used the technique to test for aluminum, beryllium, calcium, cobalt, iron, potassium, magnesium, manganese, sodium, titanium, and vanadium. No beryllium, cobalt, titanium, or vanadium occurred in any samples at detectable levels.

4.3.1 Aluminum

Aluminum, a common naturally occurring metal in the State’s groundwater, may be present in particulate form or as a solute. Current sampling procedures do not allow separate analyses of particulates and solutes. For its Secondary MCL, aluminum is subject to a range of concentrations from 50 μg/L to 200 μg/L, depending on the ability of a water system to remove the metal from water.
undergoing treatment. The EPD laboratory’s reporting level for the metal, 60 ug/L lies within the Secondary MCL range, placing any sample with detectable aluminum within the MCL range.

The metal appears to be most abundant in water samples with acidic pHs and, as a rule, is more concentrated the higher the acidity. The Miocene/Recent aquifer system, updip portions of the Cretaceous/Providence aquifer system, and updip terrigenous clastic-rich portions of the Clayton aquifer are examples. Aquifers giving mildly basic samples such as the carbonate-hosted Floridan aquifer and carbonate portions of the Valley and Ridge/Appalachian Plateau aquifers produce few sample waters containing any detectable aluminum. The metal’s abundance in bedrock waters from the Piedmont Blue Ridge aquifer system seems also low. Samples from deeper wells with more strongly basic pHs (approaching 8.00) may contain some detectable aluminum.

4.3.2 Iron and Manganese

Iron and manganese are also two more naturally occurring metals in Georgia’s groundwater. Both, like aluminum, may occur as fine particulates or as solutes. Both seem more abundant in acidic waters. Manganese also seems more abundant in waters with low dissolved oxygen contents. Sand units (e.g., the Cretaceous and updip Clayton) and shallower igneous/metamorphic bedrock give waters with the highest iron or manganese concentrations. Waters with the lowest concentrations are drawn from carbonate units (e.g., the Floridan and the carbonates in the Valley and Ridge/Appalachian Plateau province), which also usually have the higher pH waters.

4.3.3 Calcium, Magnesium, Sodium, and Potassium

Calcium is most abundant in sample waters from the Jacksonian aquifer. Sample waters from the Floridan and the Piedmont/Blue Ridge aquifer systems also contain high calcium levels. The metal could be considered least abundant in samples from the Cretaceous/Providence aquifer system. Only three, updip samples are available from the Clayton aquifer, making this lowest average calcium content hardly representative.

Magnesium appears most abundant in the Valley and Ridge/Appalachian Plateau aquifer system and least abundant in the Cretaceous/Providence system. Again, the average magnesium value for the Clayton aquifer depends on three samples and is not representative for the aquifer.

Detectable sodium is nearly ubiquitous. The metal is most abundant in waters from the Floridan and the Piedmont/Blue Ridge and least so in waters from the more updip Cretaceous.
The testing method used by the EPD laboratory to analyze for potassium is not very sensitive (reporting limit 5,000 ug/L), therefore detectable potassium was found in only seven samples from four stations – two samples from two stations in the Miocene, one sample from one station in the Floridan and four samples from one station in the Piedmont/Blue Ridge.

Kellam and Gorday (1990) observed that Ca/Mg ratios are highest in the Floridan where recharge areas are closest. Their observation also applies to the Floridan in this study, and a wide range of Ca/Mg ratios from indefinitely large (division by zero or a very small number) to 1.2 exists. However, for carbonate or carbonate-bearing aquifer media in the Valley and Ridge/Appalachian Plateau, the Jacksonian, the Claiborne, the Miocene/Surficial aquifers and aquifer systems the rule does not seem to apply. The ratios seem to cluster around 2.00 for the Valley and Ridge/Appalachian Plateau samples, and to range from 20.8 up to indefinitely large for the Jacksonian. The low number of sampling stations situated in these other aquifers or aquifer systems might cause the differences between Floridan Ca/Mg ratios and ratios for the other aquifers and aquifer systems to be apparent.

4.4 ICPMS METALS

The ICPMS method works well for most trace metals. Sample waters undergoing testing by this method, as with the samples subject to ICP testing, were unfiltered. The EPD laboratory tested for the following trace metals: chromium, nickel, copper, zinc, arsenic, selenium, molybdenum, silver, cadmium, tin, antimony, barium, thallium and lead; uranium testing was performed by the University of Georgia laboratory. Silver, cadmium, tin, antimony, and thallium remained below detection in all samples. Of the remaining metals, only lead registered any levels above the action level.

4.4.1 Chromium and Nickel

Detectable chromium occurred in one sample from one Cretaceous station and one sample from one Floridan station. Nickel occurred in one sample from one Clayton station. These metals do occur naturally occasionally in the hard metamorphic igneous rocks of the Piedmont Providence. However, in this study the chromium and nickel occurrences were in the Floridan and Clayton aquifer systems and not the Piedmont.

4.4.2 Arsenic, Selenium, Uranium, and Molybdenum

Arsenic was detected in a sample from the Floridan (quarterly well PA23). The Floridan sample came from the Gulf Trough area of Grady County, the scene of other groundwater arsenic detections, some above the Primary MCL (10 ug/L) (Donahue et al., 2012). Selenium was found in a sample from the Miocene aquifer system (Well MI10B), the Valley and Ridge aquifer system (Spring VR8) and the Floridan aquifer system (PA14A). The element may accompany uranium in
deposits formed from the reduction of oxic groundwaters. Twelve samples from three Floridan stations and one sample from one Piedmont station contained detectable molybdenum. The stations – PA23, PA28, and PA56 – are all Gulf Trough area wells. The lone sample to contain molybdenum in the Piedmont was from well BAN1A, which is a well that has had detectable uranium in the past. Like selenium, molybdenum can be associated with uranium in deposits formed through the reduction of oxic groundwaters (Turner-Peterson and Hodges, 1986). Uranium appears to be most abundant in the Piedmont/Blue Ridge, with five stations giving six samples containing detectable uranium. Uranium detections were down from previous years due to the reporting limit of the lab going from the previous 1.0 ug/L to 10 ug/L. Uranium minerals, sometimes accompanied by molybdenum and selenium minerals, can precipitate from oxic groundwaters subjected to strong reduction.

4.4.3 Copper, Lead, and Zinc

Copper, lead, and zinc detections are more numerous in acidic samples. Copper did not exceed its action level nor zinc its Secondary MCL in any samples. Out of a total of 192 samples taken for the study, 39 samples with pHs below 7.00 contained detectable amounts of at least one of these metals. In contrast, only 11 samples with basic pHs contained detectable amounts of any of these metals. Past experiences where two samples, each drawn from a different spigot, had different copper, zinc, and lead values, suggest that these metals are, at least in part, derived from plumbing. Therefore, the copper, lead, and zinc levels in the samples are not necessarily representative of those in the ambient groundwater.

4.4.4 Barium

A possible effect of the sensitivity of the testing method, barium detections occur in almost every sample. Because, perhaps, nearby barite deposits and associated mining and processing activities greatly increased the barium level in groundwater at station VR6A, a sample from that station has caused the Valley and Ridge/Appalachian Plateau samples to have one of the highest average barium levels along with samples from the Floridan and Miocene/Surficial aquifer systems. Groundwater containing excessive barium (Primary MCL of 2,000 ug/L) has not been a problem since the in-town public well field, drawing from the Floridan at Fitzgerald, Ben Hill County, closed in 1995.

4.5 CONTAMINATION OCCURENCES

According to the Safe Drinking Water Act (Public Law 93-523, section 1401, Dec. 16, 1974) a “contaminant” is any “physical, chemical, biological, or radiological substance in water” – almost anything except water itself. Some contaminants can be innocuous or even beneficial; others can be undesirable or harmful.
Modeled after limits USEPA has established concerning the quality of water offered for public consumption, the State established limits on certain contaminants in water for public use (Table 4-1). Some contaminants may endanger health, if present in sufficient concentrations. Two types of limits apply to such contaminants. The first, the Primary MCL, imposes mandatory limits applying to treated water at the point of its production. The second, the action level, sets forth mandatory limits that regulate copper and lead contents and apply to water at the point where the consumer can partake of it.

Secondary MCLs (Table 4-1) are suggested limits established for substances imparting only unpleasant qualities to water. The unpleasant qualities include bad taste and staining ability -- such as with iron and manganese -- and cosmetic effects -- such as with silver.

4.5.1 Primary MCL and Action Level Exceedances

One spring produced samples with substances that exceeded Primary MCLs or action levels (Table 4-1). Mineral spring P12A gave four samples that exceeded the Primary MCL for fluoride (4 mg/L). The spring has, in the past, regularly given samples that fall in a range from 4 mg/L to a little above 5 ug/L fluoride. The fluoride is almost certainly natural.

4.5.2 Secondary MCL Exceedances

Substances occurring in excess of Secondary MCLs (Table 4-1) consisted of manganese, iron, aluminum, sulfate, and chloride. Manganese, aluminum, and iron are common naturally occurring metals in Georgia's groundwater.

Manganese exceeded its MCL in 31 samples from 20 wells. Five of the wells were quarterly (P21, P25, P35, P37, PA34A); two gave four samples, two gave three of four samples and one gave two of four samples with excessive manganese.

The Secondary MCL for aluminum is established as a range, varying from 50 ug/L to 200 ug/L. The range is designed to accommodate varying ability of water treatment facilities at removing aluminum from treated water. This is a consequence of a tradeoff between introducing into treated water coagulants, which contain soluble aluminum, versus impaired removal of suspended aluminum-bearing contaminants. The aluminum present in waters covered by this study is naturally occurring rather than introduced. Of additional note, water in shallow wells may experience an increase in suspended matter (turbidity) during prolonged rain events, which may result in an increased aluminum value because of suspended material. Aluminum excesses, those which exceeded the 50 ug/L level (most groundwater used for public consumption lacks measureable suspended matter), were found in 17 samples from 15 wells. Aluminum excesses were the most consistent in the domestic bored Piedmont regolith well P33.
Iron equaled or exceeded its Secondary MCL in 27 samples from 24 wells. Iron is another common naturally occurring contaminant in Georgia's groundwater. Two of the wells are quarterly (P25 and P37) and well P37 had detectable iron in every quarterly sample. However, the other quarterly well (P25) only had one sample that exceeded the iron Secondary MCL.

Well P32 gave three samples with excessive sulfate and well PA9C gave a sample with excessive sulfate and excessive chloride.

4.5.3 Volatile Organic Compounds

Trihalomethanes are the most common of the VOCs detected (Table 4-2). Chloroform, the most commonly detected of the VOCs, was present in ten samples from seven stations. Bromodichloromethane was the next most common with five detections from two stations, then dibromochloromethane with four detections from one station and bromoform with one detection from one station. In groundwater, these compounds originate as by-products when halogenous disinfectants react with naturally-occurring organic matter present in the water. The disinfectants are introduced to the water through cleaning processes incident to well maintenance or through leaky check valves or foot valves allowing treated water down a well during normal operation.

One station (VR6A) gave a sample containing detectable tetrachloroethylene and 1,1-dichloroethylene. Well VR6A, an industrial process water well, is in an industrial area and is within about two miles of former and current landfills. The former landfills utilized unlined exhausted barite pits. Cressler et al. (1979) had warned of the danger of using these sorts of pits for waste disposal in the Cartersville area because of the karstic bedrock. However, the source of the VOCs at station VR6A is uncertain.

Well COU4 also gave a sample with a detection of MTBE, a fuel additive, and wells P41 and PA34A each yielded a single sample with a chloromethane detection.
### Table 4-1. Contaminant Exceedances, Calendar Year 2015.

<table>
<thead>
<tr>
<th>Station</th>
<th>Contaminant</th>
<th>MCL</th>
<th>Type Source</th>
<th>Date Sampled</th>
</tr>
</thead>
<tbody>
<tr>
<td><strong>Primary MCL and Copper/Lead Action Level Exceedances</strong></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>K12</td>
<td>Lead = 40 ug/L</td>
<td>15 ug/L</td>
<td>public well</td>
<td>1/28/15</td>
</tr>
<tr>
<td>P42</td>
<td>Lead = 23 ug/L</td>
<td>15 ug/L</td>
<td>domestic well</td>
<td>11/03/15</td>
</tr>
<tr>
<td>P12A</td>
<td>Fluoride = 4.5 mg/L</td>
<td>4 mg/L</td>
<td>mineral spring</td>
<td>02/12/15</td>
</tr>
<tr>
<td>P12A</td>
<td>Fluoride = 4.5 mg/L</td>
<td>4 mg/L</td>
<td>mineral spring</td>
<td>08/05/15</td>
</tr>
<tr>
<td>P12A</td>
<td>Fluoride = 4.4 mg/L</td>
<td>4 mg/L</td>
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<td>05/06/15</td>
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<tr>
<td>P12A</td>
<td>Fluoride = 4.7 mg/L</td>
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<tr>
<td><strong>Secondary MCL Exceedances</strong></td>
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<td></td>
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<tr>
<td>COU3</td>
<td>Manganese = 290 ug/L</td>
<td>50 ug/L</td>
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<td>08/04/15</td>
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<tr>
<td>WAS3</td>
<td>Manganese = 260 ug/L</td>
<td>50 ug/L</td>
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<td>COU4</td>
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<td>COU1</td>
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<tr>
<td>P37</td>
<td>Manganese = 150 ug/L</td>
<td>50 ug/L</td>
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<td>01/13/15</td>
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<tr>
<td>HAS1</td>
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<tr>
<td>MAD1</td>
<td>Manganese = 130 ug/L</td>
<td>50 ug/L</td>
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<td>11/17/15</td>
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<tr>
<td>P35</td>
<td>Manganese = 130 ug/L</td>
<td>50 ug/L</td>
<td>domestic well</td>
<td>01/13/15</td>
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<tr>
<td>P35</td>
<td>Manganese = 130 ug/L</td>
<td>50 ug/L</td>
<td>domestic well</td>
<td>07/22/15</td>
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<tr>
<td>P35</td>
<td>Manganese = 120 ug/L</td>
<td>50 ug/L</td>
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<tr>
<td>P35</td>
<td>Manganese = 120 ug/L</td>
<td>50 ug/L</td>
<td>domestic well</td>
<td>10/21/15</td>
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<tr>
<td>WAY1</td>
<td>Manganese = 120 ug/L</td>
<td>50 ug/L</td>
<td>public well</td>
<td>07/07/15</td>
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<tr>
<td>P37</td>
<td>Manganese = 110 ug/L</td>
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<td>07/22/15</td>
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<tr>
<td>PA34A</td>
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<td>12/01/15</td>
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<tr>
<td>PA34A</td>
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<td>50 ug/L</td>
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<td>06/23/15</td>
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<td>PA34A</td>
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4-11
<table>
<thead>
<tr>
<th>Station</th>
<th>Contaminant</th>
<th>MCL</th>
<th>Type</th>
<th>Source</th>
<th>Date Sampled</th>
</tr>
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<tr>
<td>SUM2</td>
<td>Manganese = 82 ug/L</td>
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<td>Manganese = 77 ug/L</td>
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<td>P20</td>
<td>Manganese = 76 ug/L</td>
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<td>10/07/15</td>
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<td>JEF1</td>
<td>Manganese = 71 ug/L</td>
<td>50 ug/L</td>
<td>domestic well</td>
<td>domestic well</td>
<td>10/06/15</td>
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<tr>
<td>MI10B</td>
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<td>50 ug/L</td>
<td>domestic well</td>
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<tr>
<td>P25</td>
<td>Manganese = 63 ug/L</td>
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<td>PA18</td>
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<tr>
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<tr>
<td>CL4A</td>
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<td>P21</td>
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<td>FRA1</td>
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<tr>
<td>CL8</td>
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<td>50 ug/L</td>
<td>public well</td>
<td>public well</td>
<td>01/28/15</td>
</tr>
<tr>
<td>P21</td>
<td>Manganese = 53 ug/L</td>
<td>50 ug/L</td>
<td>public well</td>
<td>public well</td>
<td>02/11/15</td>
</tr>
<tr>
<td>SUM2</td>
<td>Aluminum=1,100ug/L</td>
<td>50-200 ug/L</td>
<td>public well</td>
<td>public well</td>
<td>01/28/15</td>
</tr>
<tr>
<td>K12</td>
<td>Aluminum = 340 ug/L</td>
<td>50-200 ug/L</td>
<td>public well</td>
<td>public well</td>
<td>01/28/15</td>
</tr>
<tr>
<td>K9A</td>
<td>Aluminum = 260 ug/L</td>
<td>50-200 ug/L</td>
<td>public well</td>
<td>public well</td>
<td>01/14/15</td>
</tr>
<tr>
<td>CHT3</td>
<td>Aluminum = 200 ug/L</td>
<td>50-200 ug/L</td>
<td>public well</td>
<td>public well</td>
<td>11/18/15</td>
</tr>
<tr>
<td>MI2A</td>
<td>Aluminum = 190 ug/L</td>
<td>50-200 ug/L</td>
<td>domestic well</td>
<td>domestic well</td>
<td>08/26/15</td>
</tr>
<tr>
<td>GLA1</td>
<td>Aluminum = 150 ug/L</td>
<td>50-200 ug/L</td>
<td>public well</td>
<td>public well</td>
<td>10/06/15</td>
</tr>
<tr>
<td>BUR2</td>
<td>Aluminum = 130 ug/L</td>
<td>50-200 ug/L</td>
<td>public well</td>
<td>public well</td>
<td>02/11/15</td>
</tr>
<tr>
<td>PA38</td>
<td>Aluminum = 130 ug/L</td>
<td>50-200 ug/L</td>
<td>public well</td>
<td>public well</td>
<td>06/23/15</td>
</tr>
<tr>
<td>P33</td>
<td>Aluminum = 95 ug/L</td>
<td>50-200 ug/L</td>
<td>domestic well</td>
<td>domestic well</td>
<td>01/13/15</td>
</tr>
<tr>
<td>P33</td>
<td>Aluminum = 83 ug/L</td>
<td>50-200 ug/L</td>
<td>domestic well</td>
<td>domestic well</td>
<td>10/21/15</td>
</tr>
<tr>
<td>PA36</td>
<td>Aluminum = 83 ug/L</td>
<td>50-200 ug/L</td>
<td>public well</td>
<td>public well</td>
<td>09/09/15</td>
</tr>
<tr>
<td>Station</td>
<td>Contaminant</td>
<td>MCL</td>
<td>Type Source</td>
<td>Date Sampled</td>
<td></td>
</tr>
<tr>
<td>---------</td>
<td>------------</td>
<td>-----</td>
<td>-------------</td>
<td>--------------</td>
<td></td>
</tr>
<tr>
<td>P33</td>
<td>Aluminum = 80 ug/L</td>
<td>50-200 ug/L</td>
<td>domestic well</td>
<td>04/08/15</td>
<td></td>
</tr>
<tr>
<td>GLY4</td>
<td>Aluminum = 78 ug/L</td>
<td>50-200 ug/L</td>
<td>public well</td>
<td>09/10/15</td>
<td></td>
</tr>
<tr>
<td>P25</td>
<td>Aluminum = 69 ug/L</td>
<td>50-200 ug/L</td>
<td>public well</td>
<td>11/04/15</td>
<td></td>
</tr>
<tr>
<td>CT8</td>
<td>Aluminum = 69 ug/L</td>
<td>50-200 ug/L</td>
<td>domestic well</td>
<td>09/23/15</td>
<td></td>
</tr>
<tr>
<td>PA34A</td>
<td>Aluminum = 64 ug/L</td>
<td>50-200 ug/L</td>
<td>public well</td>
<td>09/09/15</td>
<td></td>
</tr>
<tr>
<td>FRA1</td>
<td>Iron = 7,400 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>07/22/15</td>
<td></td>
</tr>
<tr>
<td>P41</td>
<td>Iron = 3,200 ug/L</td>
<td>300 ug/L</td>
<td>domestic well</td>
<td>11/06/15</td>
<td></td>
</tr>
<tr>
<td>P37</td>
<td>Iron = 3,000 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>01/13/15</td>
<td></td>
</tr>
<tr>
<td>BUR2</td>
<td>Iron = 2,200 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>02/11/15</td>
<td></td>
</tr>
<tr>
<td>CL4A</td>
<td>Iron = 2,100 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>01/28/15</td>
<td></td>
</tr>
<tr>
<td>MI10B</td>
<td>Iron = 1,600 ug/L</td>
<td>300 ug/L</td>
<td>domestic well</td>
<td>08/26/15</td>
<td></td>
</tr>
<tr>
<td>CHT1</td>
<td>Iron = 1,500 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>11/18/15</td>
<td></td>
</tr>
<tr>
<td>COU3</td>
<td>Iron = 1,400 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>08/04/15</td>
<td></td>
</tr>
<tr>
<td>P37</td>
<td>Iron = 1,300 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>04/08/15</td>
<td></td>
</tr>
<tr>
<td>P37</td>
<td>Iron = 1,200 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>07/22/15</td>
<td></td>
</tr>
<tr>
<td>STW1</td>
<td>Iron = 1,200 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>09/23/15</td>
<td></td>
</tr>
<tr>
<td>GLY4</td>
<td>Iron = 1,200 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>09/10/15</td>
<td></td>
</tr>
<tr>
<td>STW2</td>
<td>Iron = 1,100 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>09/23/15</td>
<td></td>
</tr>
<tr>
<td>MAC1</td>
<td>Iron = 1,000 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>01/29/15</td>
<td></td>
</tr>
<tr>
<td>COU1</td>
<td>Iron = 1,000 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>08/04/15</td>
<td></td>
</tr>
<tr>
<td>PA9C</td>
<td>Iron = 1,000 ug/L</td>
<td>300 ug/L</td>
<td>former test</td>
<td>09/10/15</td>
<td></td>
</tr>
<tr>
<td>K3</td>
<td>Iron = 960 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>10/06/15</td>
<td></td>
</tr>
<tr>
<td>GLY2</td>
<td>Iron = 800 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>09/10/15</td>
<td></td>
</tr>
<tr>
<td>P25</td>
<td>Iron = 670 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>02/12/15</td>
<td></td>
</tr>
</tbody>
</table>

*Secondary MCL Exceedances Continued*
<table>
<thead>
<tr>
<th>Station</th>
<th>Contaminant</th>
<th>MCL</th>
<th>Type Source</th>
<th>Date Sampled</th>
</tr>
</thead>
<tbody>
<tr>
<td>MAD1</td>
<td>Iron = 660 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>11/17/15</td>
</tr>
<tr>
<td>SUM2</td>
<td>Iron = 610 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>01/28/15</td>
</tr>
<tr>
<td>CL8</td>
<td>Iron = 570 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>01/28/15</td>
</tr>
<tr>
<td>K21</td>
<td>Iron = 480 ug/L</td>
<td>300 ug/L</td>
<td>domestic well</td>
<td>02/25/15</td>
</tr>
<tr>
<td>K12</td>
<td>Iron = 450 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>01/28/15</td>
</tr>
<tr>
<td>GLA1</td>
<td>Iron = 400 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>10/06/15</td>
</tr>
<tr>
<td>P37</td>
<td>Iron = 330 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>10/21/15</td>
</tr>
<tr>
<td>CHT3</td>
<td>Iron = 320 ug/L</td>
<td>300 ug/L</td>
<td>public well</td>
<td>11/18/15</td>
</tr>
<tr>
<td>P32</td>
<td>Sulfate = 340 mg/L</td>
<td>250 mg/L</td>
<td>domestic well</td>
<td>04/08/15</td>
</tr>
<tr>
<td>P32</td>
<td>Sulfate = 340 mg/L</td>
<td>250 mg/L</td>
<td>domestic well</td>
<td>10/21/15</td>
</tr>
<tr>
<td>PA9C</td>
<td>Sulfate = 280 mg/L</td>
<td>250 mg/L</td>
<td>former test</td>
<td>09/10/15</td>
</tr>
<tr>
<td>P32</td>
<td>Sulfate = 270 mg/L</td>
<td>250 mg/L</td>
<td>domestic well</td>
<td>01/13/15</td>
</tr>
<tr>
<td>PA9C</td>
<td>Chloride = 820 mg/L</td>
<td>250 mg/L</td>
<td>former test</td>
<td>09/10/15</td>
</tr>
</tbody>
</table>

(The alphabetic prefix in a station number indicates the aquifer/aquifer system tapped: CL=Claiborne, J=Jacksonian, K=Cretaceous, P=Piedmont/Blue Ridge, PA=Floridan, CT=Clayton, VR=Valley and Ridge, M=Miocene)
<table>
<thead>
<tr>
<th>Station</th>
<th>Constituents</th>
<th>Primary MCL</th>
<th>Type Source</th>
<th>Date Sampled</th>
</tr>
</thead>
<tbody>
<tr>
<td>GWN-UPS1</td>
<td>chloroform = 1.40 ug/L</td>
<td>See note</td>
<td>public</td>
<td>03/12/15</td>
</tr>
<tr>
<td>GWN-SUM1</td>
<td>chloroform = 0.78 ug/L</td>
<td>See note</td>
<td>public</td>
<td>01/28/15</td>
</tr>
<tr>
<td>GWN-K7</td>
<td>chloroform = 2.00 ug/L</td>
<td>See note</td>
<td>public</td>
<td>02/25/15</td>
</tr>
<tr>
<td>GWN-J4</td>
<td>chloroform = 0.57 ug/L</td>
<td>See note</td>
<td>public</td>
<td>02/26/15</td>
</tr>
<tr>
<td>GWN-P41</td>
<td>chloromethane = 0.66 ug/L</td>
<td>See note</td>
<td>domestic</td>
<td>11/06/15</td>
</tr>
<tr>
<td>GWN-PA34A</td>
<td>chloromethane = 0.79 ug/L</td>
<td>See note</td>
<td>public</td>
<td>12/01/15</td>
</tr>
<tr>
<td>GWN-PA17</td>
<td>chloroform = 0.64 ug/L</td>
<td>See note</td>
<td>public</td>
<td>03/10/15</td>
</tr>
<tr>
<td></td>
<td>bromodichloromethane = 0.52 ug/L</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>GWN-PA23</td>
<td>chloroform = 1.20 ug/L</td>
<td>See note</td>
<td>public</td>
<td>01/14/15</td>
</tr>
<tr>
<td></td>
<td>bromodichloromethane = 0.91 ug/L</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>GWN-PA23</td>
<td>chloroform = 5.90 ug/L</td>
<td>See note</td>
<td>public</td>
<td>04/21/15</td>
</tr>
<tr>
<td></td>
<td>bromodichloromethane = 3.60 ug/L</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td></td>
<td>dibromochloromethane = 2.70 ug/L</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td></td>
<td>bromoform = 0.57 ug/L</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>GWN-PA23</td>
<td>chloroform = 2.70 ug/L</td>
<td>See note</td>
<td>public</td>
<td>07/28/15</td>
</tr>
<tr>
<td></td>
<td>bromodichloromethane = 1.40 ug/L</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td></td>
<td>dibromochloromethane = 1.20 ug/L</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>GWN-PA23</td>
<td>chloroform = 5.00 ug/L</td>
<td>See note</td>
<td>public</td>
<td>10/20/15</td>
</tr>
<tr>
<td></td>
<td>bromodichloromethane = 2.20 ug/L</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td></td>
<td>dibromochloromethane = 1.70 ug/L</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>GWN-COU4</td>
<td>chloroform = 0.64 ug/L</td>
<td>See note</td>
<td>public</td>
<td>11/03/15</td>
</tr>
<tr>
<td></td>
<td>MTBE = 1.10 ug/L</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>GWN-VR6A</td>
<td>1,1 dichloroethylene = 2.10 ug/L</td>
<td>7 ug/L</td>
<td>public</td>
<td>05/20/15</td>
</tr>
<tr>
<td></td>
<td>tetrachloroethylene = 2.30 ug/L</td>
<td>5 ug/L</td>
<td></td>
<td></td>
</tr>
</tbody>
</table>
4.6 GENERAL QUALITY

A review of the analyses of the water samples collected during calendar year 2015 indicates that the chemical quality of groundwater sampled for most of the Groundwater Monitoring Network stations is quite good.

However, as mentioned in Chapter 1, areas of elevated risk for low-quality groundwater exist:

1) Valley and Ridge/Appalachian Plateau Province – surface influence;
2) Piedmont/Blue Ridge Province – in areas excluding the eastern metavolcanic terranes - uranium:
3) Coastal Plain agricultural areas – high nitrate/nitrite;
4) Coastal Plain, Dougherty Plain – surface influence;
5) Coastal Plain, Gulf Trough – high total dissolved solids, especially sulfate – high radionuclides, high barium, high arsenic;
6) Coastal Plain, Atlantic coast area – saline water influx.
CHAPTER 5 LIST OF REFERENCES

Applied Coastal Research Laboratory, Georgia Southern University, 2002, Gulf Trough and Satilla Line Data Analysis, Georgia Geologic Survey Project Report 48, 14 p., 1 pl.


LABORATORY AND STATION DATA

Tables A-1 through A-8 list the values for both laboratory parameters and field parameters for each well or spring. The following abbreviations are used on these tables:

<table>
<thead>
<tr>
<th>Parameters and Units of Measure</th>
<th>ND</th>
<th>= not detected</th>
</tr>
</thead>
<tbody>
<tr>
<td>Cl = chloride</td>
<td>NG</td>
<td>= not given</td>
</tr>
<tr>
<td>cond. = conductivity</td>
<td>NOx</td>
<td>= nitrate/nitrite</td>
</tr>
<tr>
<td>diss O2 = dissolved oxygen</td>
<td>P</td>
<td>= total phosphorus</td>
</tr>
<tr>
<td>F = fluoride</td>
<td>SO4</td>
<td>= sulfate</td>
</tr>
<tr>
<td>ICP = inductively coupled</td>
<td>Temp.</td>
<td>= temperature</td>
</tr>
<tr>
<td>plasma (emission) spectroscopy</td>
<td>ug/L</td>
<td>= micrograms per liter</td>
</tr>
<tr>
<td>mg/L = milligrams per liter</td>
<td>uS/cm</td>
<td>= microSiemenses per centimeter</td>
</tr>
<tr>
<td>mgN/L = milligrams per liter as nitrogen</td>
<td>VOC</td>
<td>= volatile organic compound</td>
</tr>
<tr>
<td>NA = not available; not</td>
<td></td>
<td></td>
</tr>
<tr>
<td>analyzed</td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

Volatile Organic Compounds

| 1,1dce = 1,1-dichloroethylene | mdcb | = m-dichlorobenzene |
| bdcm = bromodichloromethane    | odcb | = o-dichlorobenzene |
| dbcm = dibromochloromethane    | pdcb | = p-dichlorobenzene |
| pce = tetrachloroethylene      | tbm  | = bromoform        |
| cb = chlorobenzene             | tcm  | = chloroform       |
| MTBE = methyl tert-butyl ether |     |                  |

Table A-9 gives the reporting limits for the various analytes. The abbreviations used for Tables A-1 through A-8 also apply to Table A-9.
Table A-1. Groundwater Quality Analyses for Cretaceous/Providence Stations.
Part A: Station Identification, Date of Sampling, Field Parameters, VOCs, Anions, and Non-Metals.

<table>
<thead>
<tr>
<th>Station No.</th>
<th>Well Name</th>
<th>Wet Depth (feet)</th>
<th>Casing Depth (feet)</th>
<th>Well Size (Inches)</th>
<th>Date sampled</th>
<th>pH</th>
<th>cond (uSi/cm)</th>
<th>disCl (mg/L)</th>
<th>Temp (°C)</th>
<th>VOCS (ppb)</th>
<th>Cl (mg/L)</th>
<th>SO4 (mg/L)</th>
<th>NOX (mg N/L)</th>
<th>P (mg/L)</th>
</tr>
</thead>
<tbody>
<tr>
<td>GWN-K3</td>
<td>Sandersville Well #7B</td>
<td>697</td>
<td>NG</td>
<td>NG</td>
<td>10/05/15</td>
<td>6.27</td>
<td>149</td>
<td>7.90</td>
<td>20.39</td>
<td>ND</td>
<td>ND</td>
<td>ND</td>
<td>0.49</td>
<td>0.10</td>
</tr>
<tr>
<td>GWN-K6</td>
<td>KaMin Well #6</td>
<td>400</td>
<td>NG</td>
<td>NG</td>
<td>06/10/15</td>
<td>5.27</td>
<td>45</td>
<td>3.14</td>
<td>20.08</td>
<td>ND</td>
<td>ND</td>
<td>ND</td>
<td>0.02</td>
<td>ND</td>
</tr>
<tr>
<td>GWN-K7</td>
<td>Jones County #4</td>
<td>128</td>
<td>NG</td>
<td>NG</td>
<td>02/25/15</td>
<td>5.12</td>
<td>37</td>
<td>6.00</td>
<td>18.03</td>
<td>18.00</td>
<td>ND</td>
<td>ND</td>
<td>0.41</td>
<td>ND</td>
</tr>
<tr>
<td>GWN-K9A</td>
<td>Marshallville Well #2</td>
<td>550</td>
<td>NG</td>
<td>NG</td>
<td>01/14/15</td>
<td>3.99</td>
<td>49</td>
<td>NA</td>
<td>18.63</td>
<td>ND</td>
<td>ND</td>
<td>ND</td>
<td>0.04</td>
<td>ND</td>
</tr>
<tr>
<td>GWN-K10B</td>
<td>Fort Valley Well #6</td>
<td>600</td>
<td>NG</td>
<td>NG</td>
<td>03/12/15</td>
<td>4.85</td>
<td>20</td>
<td>10.70</td>
<td>18.27</td>
<td>ND</td>
<td>ND</td>
<td>ND</td>
<td>0.74</td>
<td>ND</td>
</tr>
<tr>
<td>GWN-K11A</td>
<td>Warner Robins Well #2</td>
<td>540</td>
<td>NG</td>
<td>NG</td>
<td>01/28/15</td>
<td>4.76</td>
<td>28</td>
<td>8.61</td>
<td>18.96</td>
<td>ND</td>
<td>ND</td>
<td>ND</td>
<td>1.0</td>
<td>ND</td>
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<tr>
<td>GWN-K112</td>
<td>Perry/Holiday Inn Well</td>
<td>550</td>
<td>NG</td>
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Table A-1: Groundwater Quality Analyses for Cretaceous Stations.

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**Note:** The table above contains a list of stations with their respective counties and analytical results for various metals. The data is presented in a tabular format with columns indicating different stations and rows showing the analytical results for each station.
Table A-1. Groundwater Quality Analyses for Cretaceous/Providence Stations.  
Part A: Station Identification, Date of Sampling, Field Parameters, VOCs, Anions, and Non-Metals.

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<th>Casing Depth feet</th>
<th>Well Size Inches</th>
<th>Date sampled</th>
<th>pH</th>
<th>cond. µS/cm</th>
<th>Diss O2 mg/L</th>
<th>Temp °C</th>
<th>VOCs ug/L</th>
<th>CI mg/L</th>
<th>Sulfate mg/L</th>
<th>NOx mg/L</th>
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Table A-2. Groundwater Quality Analyses for Clayton Stations.
Part A: Station Identification, Date of Sampling, Field Parameters, VOCs, Anions, and Non-Metals.

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<th>cond. µS/cm</th>
<th>diss O2 mg/L</th>
<th>Temp °C</th>
<th>VOCS ug/L</th>
<th>Cl mg/L</th>
<th>SO4 mg/L</th>
<th>NOx mg N/L</th>
<th>P mg/l</th>
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Table A-2: Groundwater Quality Analyses for Clayton Stations.
Table A-3. Groundwater Quality Analyses for Claiborne Stations.
Part A: Station Identification, Date of Sampling, Field Parameters, VOCs, Anions, and Non-Metals.

<table>
<thead>
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<th>Station No.</th>
<th>Well Name</th>
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<th>Casing Depth feet</th>
<th>Well Size Inches</th>
<th>Date sampled</th>
<th>pH</th>
<th>cond. uS/cm</th>
<th>Diss O2 mg/L</th>
<th>Temp °C</th>
<th>VOCs ug/L</th>
<th>Cl mg/L</th>
<th>SO4 mg/L</th>
<th>NOx mg Nit.</th>
<th>P mg/L</th>
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Table A-3. Groundwater Quality Analyses for Claiborne Stations.
Part B: Metals.

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<th>Tin</th>
<th>Antimony</th>
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<th>Thallium</th>
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Table A-4. Groundwater Quality Analyses for Jacksonian Stations.
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## Table A-5, Continued. Groundwater Quality Analyses for Floridan Stations.

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Table A-5, Continued. Groundwater Quality Analyses for Floridan Stations.
Part A: Station Identification, Date of Sampling, Field Parameters, VOCs, Anions, and Non-Metals.

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<th>Temp °C</th>
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Table A-6. Groundwater Quality Analyses for Miocene Stations.  
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### Table A-6. Groundwater Quality Analyses for Miocene Stations.
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### Table A-7. Groundwater Quality Analyses for Piedmont-Blue Ridge Stations.

**Part A: Station Identification, Date of Sampling, Field Parameters, VOCs, Anions, and Non-Metals.**

| Station No. | Well Name | County | Well Depth (feet) | Casing Depth (feet) | Well Size (inches) | Date Sampled | pH | cond. (μS/cm) | diss CO2 (mg/L) | Temp (°C) | VOCS (μg/L) | F (mg/L) | Cl (mg/L) | SO4 (mg/L) | NO3 (mg/L) | P (mg/L) |
|-------------|-----------|--------|------------------|--------------------|-------------------|--------------|----|--------------|----------------|-----------|-------------|----------|----------|------------|------------|---------|----------|
| GWN-P1A     | Luthersville Well #3 | Meriwether | 185              | NG                | NG                | 06/24/15     | 6.08 | 87           | 7.48            | 17.60     | ND          | NA       | ND       | ND         | 1.10       | 0.08    |
| GWN-P5      | Flowery Branch Well #1 | Hall    | 240              | NG                | NG                | 10/07/15     | 6.70 | 179          | 0.78            | 18.38     | ND          | NA       | ND       | ND         | 1.10       | 0.03    |
| GWN-P12A    | Indian Spring | Butts   | 0                | NG                | NG                | 02/12/15     | 7.10 | 267          | NA              | 16.08     | ND          | 4.5       | 10       | 24         | ND         | 0.02    |
|             |           |         |                  |                    |                   | 05/06/15     | 7.36 | 262          | NA              | 17.88     | ND          | 4.4       | 10       | 24         | ND         | 0.02    |
|             |           |         |                  |                    |                   | 06/05/15     | 7.40 | 265          | NA              | 19.93     | ND          | 4.5       | 10       | 24         | ND         | 0.02    |
|             |           |         |                  |                    |                   | 11/04/15     | 7.44 | 270          | NA              | 18.09     | ND          | 4.7       | 11       | 25         | ND         | 0.03    |
| GWN-P20     | Suwanee #1 | Gwinnett | 600              | NG                | NG                | 10/07/15     | 7.69 | 347          | 4.13            | 17.58     | ND          | NA       | ND       | 13         | 0.23       | ND      |
| GWN-P21     | Gray/Bragg Well | Jones | 405              | NG                | NG                | 02/11/15     | 6.80 | 322          | 10.12           | 19.14     | ND          | NA       | ND       | 26         | 0.09       | 0.03    |
|             |           |         |                  |                    |                   | 05/06/15     | 6.73 | 303          | 8.88            | 19.28     | ND          | NA       | ND       | 27         | 0.10       | 0.03    |
|             |           |         |                  |                    |                   | 08/05/15     | 6.80 | 311          | 6.50            | 19.52     | ND          | NA       | ND       | 26         | 0.09       | 0.03    |
|             |           |         |                  |                    |                   | 11/04/15     | 6.93 | 307          | 7.04            | 19.24     | ND          | NA       | ND       | 31         | 0.33       | 0.03    |
| GWN-P22     | Rahbar Well | Fulton | 200              | NG                | NG                | 04/22/15     | 4.67 | 35           | 4.67            | 17.18     | ND          | NA       | ND       | ND         | 0.92       | ND      |
| GWN-P23     | Indian Springs State Park New Main Well | Butts | NG | NG | 02/12/15 | 6.60 | 153 | 2.61 | 18.89 | ND | 1.1 | ND | ND | 0.21 | 0.07 |
|             |           |         |                  |                    |                   | 05/06/15     | 6.52 | 145          | 2.40            | 18.20     | ND          | 1.1       | ND       | 0.23       | 0.07       | ND      |
|             |           |         |                  |                    |                   | 08/05/15     | 6.62 | 152          | 2.40            | 18.30     | ND          | 1.2       | ND       | 0.19       | 0.07       | ND      |
|             |           |         |                  |                    |                   | 11/04/15     | 6.64 | 151          | 2.27            | 18.16     | ND          | 1.2       | ND       | 0.21       | 0.07       | ND      |
| GWN-P24     | The Gates Well #1 | Coweta | 705              | NG                | NG                | 04/23/15     | 7.11 | 268          | 2.04            | 19.76     | ND          | NA       | ND       | 11         | 0.26       | 0.04    |
| GWN-P25     | Jarrell Plantation Staff House Well | Jones | NG | NG | 02/12/15 | 6.12 | 170 | 4.10 | 18.78 | ND | NA | ND | ND | 0.26 | 0.18 |
|             |           |         |                  |                    |                   | 05/06/15     | 6.09 | 186          | 3.88            | 18.31     | ND          | NA       | ND       | 0.26       | 0.12       | ND      |
|             |           |         |                  |                    |                   | 08/05/15     | 6.26 | 185          | 8.51            | 18.53     | ND          | NA       | ND       | 0.20       | 0.11       | ND      |
|             |           |         |                  |                    |                   | 11/04/15     | 6.28 | 184          | 3.85            | 18.36     | ND          | NA       | ND       | 0.21       | 0.12       | ND      |
| GWN-P28     | Willow Court Well | Coweta | NG | NG | 04/23/15 | 5.88 | 138 | 4.02 | 18.11 | ND | NA | ND | ND | 1.7 | 0.07 |
| GWN-P30     | Fizer House Well | Lincoln | 220              | NG                | NG                | 05/05/15     | 6.87 | 461          | 2.26            | 19.04     | ND          | NA       | 23       | 23         | 3.1        | 0.04    |
| GWN-P32     | Cecchin Deep Well | Elbert | 400              | NG                | NG                | 01/13/15     | 7.80 | 737          | 0.81            | 15.26     | ND          | NA       | ND       | 270        | ND         | ND      |
|             |           |         |                  |                    |                   | 04/08/15     | 7.76 | 964          | 0.83            | 17.75     | ND          | NA       | ND       | 340        | ND         | ND      |
|             |           |         |                  |                    |                   | 07/22/15     | 7.87 | 677          | 7.90            | 20.42     | ND          | NA       | ND       | 210        | ND         | ND      |
|             |           |         |                  |                    |                   | 10/21/15     | 7.79 | 895          | 0.82            | 17.54     | ND          | NA       | ND       | 340        | ND         | ND      |
| GWN-P33     | Cecchin Bored Well | Elbert | 47               | NG                | NG                | 01/13/15     | 6.44 | 109          | 6.10            | 16.88     | ND          | NA       | ND       | 0.92       | 0.03       | ND      |
|             |           |         |                  |                    |                   | 04/08/15     | 6.00 | 97           | 8.10            | 17.42     | ND          | NA       | ND       | 1.00       | 0.02       | ND      |
|             |           |         |                  |                    |                   | 07/22/15     | 6.15 | 106          | 8.87            | 17.95     | ND          | NA       | ND       | 2.00       | 0.02       | ND      |
|             |           |         |                  |                    |                   | 10/21/15     | 6.34 | 100          | 6.80            | 17.80     | ND          | NA       | ND       | 2.50       | 0.02       | ND      |
Table A-7. Groundwater Quality Analyses for Piedmont-Blue Ridge Stations.

Part B: Metals.

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A.21
Table A-7 Continued. Groundwater Quality Analyses for Piedmont-Blue Ridge Stations.
Part A: Station Identification, Date of Sampling, Field Parameters, VOCs, Anions, and Non-Metals.

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Table A-7 Continued. Groundwater Quality Analyses for Piedmont-Blue Ridge Stations.  
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<th>dias O2 mg/L</th>
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<td>ND</td>
<td>ND</td>
<td>ND</td>
<td>ND</td>
<td>ND</td>
<td>4.6</td>
<td>ND</td>
<td>ND</td>
<td>ND</td>
<td>ND</td>
<td>33,000</td>
<td>ND</td>
<td>ND</td>
<td>ND</td>
<td>ND</td>
<td>ND</td>
<td>15,000</td>
<td>ND</td>
<td>ND</td>
<td>ND</td>
<td>ND</td>
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<tr>
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<td></td>
</tr>
</tbody>
</table>

A-27
Table A-9. Analytes, EPA Analytical Methods, and Reporting Limits.

<table>
<thead>
<tr>
<th>Analyte</th>
<th>Reporting Limit/ EPA Method</th>
<th>Analyte</th>
<th>Reporting Limit/ EPA Method</th>
</tr>
</thead>
<tbody>
<tr>
<td>Vinyl Chloride</td>
<td>0.5 µg/L / 524.2</td>
<td>Dichlorodifluoromethane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>1,1-Dichloroethylene</td>
<td>0.5 µg/L / 524.2</td>
<td>Chloromethane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>Dichloromethane</td>
<td>0.5 µg/L / 524.2</td>
<td>Bromomethane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>Trans-1,2-Dichloroethylene</td>
<td>0.5 µg/L / 524.2</td>
<td>Chloroethane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>Cis-1,2-Dichloroethylene</td>
<td>0.5 µg/L / 524.2</td>
<td>Fluorotrichloromethane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>1,1,1-Trichloroethane</td>
<td>0.5 µg/L / 524.2</td>
<td>1,1-Dichloroethane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>Carbon Tetrachloride</td>
<td>0.5 µg/L / 524.2</td>
<td>2,2-Dichloropropane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>Benzene</td>
<td>0.5 µg/L / 524.2</td>
<td>Bromochloromethane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>1,2-Dichloroethane</td>
<td>0.5 µg/L / 524.2</td>
<td>Chloroform</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>Trichloroethylene</td>
<td>0.5 µg/L / 524.2</td>
<td>1,1-Dichloropropene</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>1,2-Dichloropropane</td>
<td>0.5 µg/L / 524.2</td>
<td>Dibromomethane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>Toluene</td>
<td>0.5 µg/L / 524.2</td>
<td>Bromodichloromethane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>1,1,2-Trichloroethane</td>
<td>0.5 µg/L / 524.2</td>
<td>Cis-1,3-Dichloropropene</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>Tetrachloroethylene</td>
<td>0.5 µg/L / 524.2</td>
<td>Trans-1,3-Dichloropropene</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>Chlorobenzene</td>
<td>0.5 µg/L / 524.2</td>
<td>1,3-Dichloropropane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>Ethylbenzene</td>
<td>0.5 µg/L / 524.2</td>
<td>Chlorodibromomethane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>Total Xylenes</td>
<td>0.5 µg/L / 524.2</td>
<td>1,2-Dibromoethane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>Styrene</td>
<td>0.5 µg/L / 524.2</td>
<td>1,1,1,2-Tetrachloroethane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>p-Dichlorobenzene</td>
<td>0.5 µg/L / 524.2</td>
<td>Bromoform</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>o-Dichlorobenzene</td>
<td>0.5 µg/L / 524.2</td>
<td>Isopropylbenzene</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>1,2,4-Trichlorobenzene</td>
<td>0.5 µg/L / 524.2</td>
<td>1,1,2,2-Tetrachloroethane</td>
<td>0.5 µg/L / 524.2</td>
</tr>
<tr>
<td>Analyte</td>
<td>Reporting Limit/ EPA Method</td>
<td>Analyte</td>
<td>Reporting Limit/ EPA Method</td>
</tr>
<tr>
<td>----------------------------------------------</td>
<td>-----------------------------</td>
<td>----------------------------------</td>
<td>-----------------------------</td>
</tr>
<tr>
<td>Bromobenzene</td>
<td>0.5 ug/L / 524.2</td>
<td>Total Phosphorus</td>
<td>0.02 mg/L / 365.1</td>
</tr>
<tr>
<td>1,2,3-Trichloro-propane</td>
<td>0.5 ug/L / 524.2</td>
<td>Fluoride</td>
<td>0.20 mg/L / 300.0</td>
</tr>
<tr>
<td>n-Propylbenzene</td>
<td>0.5 ug/L / 524.2</td>
<td>Silver</td>
<td>10 ug/L / 200.7 (ICP)</td>
</tr>
<tr>
<td>o-Chlorotoluene</td>
<td>0.5 ug/L / 524.2</td>
<td>Aluminum</td>
<td>60 ug/L / 200.7</td>
</tr>
<tr>
<td>1,3,5-Trimethyl-benzene</td>
<td>0.5 ug/L / 524.2</td>
<td>Arsenic</td>
<td>80 ug/L / 200.7</td>
</tr>
<tr>
<td>p-Chlorotoluene</td>
<td>0.5 ug/L / 524.2</td>
<td>Barium</td>
<td>10 ug/L / 200.7</td>
</tr>
<tr>
<td>Tert-Butylbenzene</td>
<td>0.5 ug/L / 524.2</td>
<td>Beryllium</td>
<td>10 ug/L / 200.7</td>
</tr>
<tr>
<td>1,2,4-Trimethyl-benzene</td>
<td>0.5 ug/L / 524.2</td>
<td>Calcium</td>
<td>1000 ug/L / 200.7</td>
</tr>
<tr>
<td>Sec-Butylbenzene</td>
<td>0.5 ug/L / 524.2</td>
<td>Cobalt</td>
<td>10 ug/L / 200.7</td>
</tr>
<tr>
<td>p-Isopropyltoluene</td>
<td>0.5 ug/L / 524.2</td>
<td>Chromium</td>
<td>20 ug/L / 200.7</td>
</tr>
<tr>
<td>m-Dichlorobenzene</td>
<td>0.5 ug/L / 524.2</td>
<td>Copper</td>
<td>20 ug/L / 200.7</td>
</tr>
<tr>
<td>n-Butylbenzene</td>
<td>0.5 ug/L / 524.2</td>
<td>Iron</td>
<td>20 ug/L / 200.7</td>
</tr>
<tr>
<td>1,2-Dibromo-3-chloropropane</td>
<td>0.5 ug/L / 524.2</td>
<td>Potassium</td>
<td>5000 ug/L / 200.7</td>
</tr>
<tr>
<td>Hexachlorobutadiene</td>
<td>0.5 ug/L / 524.2</td>
<td>Magnesium</td>
<td>1000 ug/L / 200.7</td>
</tr>
<tr>
<td>Naphthalene</td>
<td>0.5 ug/L / 524.2</td>
<td>Manganese</td>
<td>10 ug/L / 200.7</td>
</tr>
<tr>
<td>1,2,3-Trichlorobenzene</td>
<td>0.5 ug/L / 524.2</td>
<td>Sodium</td>
<td>1000 ug/L / 200.7</td>
</tr>
<tr>
<td>Methyl-tert-butyl ether (MTBE)</td>
<td>0.5 ug/L / 524.2</td>
<td>Nickel</td>
<td>20 ug/L / 200.7</td>
</tr>
<tr>
<td>Chloride</td>
<td>10 mg/L / 300.0</td>
<td>Lead</td>
<td>90 ug/L / 200.7</td>
</tr>
<tr>
<td>Sulfate*</td>
<td>10 mg/L / 300.0</td>
<td>Antimony</td>
<td>120 ug/L / 200.7</td>
</tr>
<tr>
<td>Nitrate/nitrite*</td>
<td>0.02 mg/L as Nitrogen / 353.2</td>
<td>Selenium</td>
<td>190 ug/L / 200.7</td>
</tr>
</tbody>
</table>
Table A-9, Continued. Analytes, EPA Analytical Methods, and Reporting Limits.

<table>
<thead>
<tr>
<th>Analyte</th>
<th>Reporting Limit/ EPA Method</th>
<th>Analyte</th>
<th>Reporting Limit/ EPA Method</th>
</tr>
</thead>
<tbody>
<tr>
<td>Titanium</td>
<td>10 ug/L / 200.7</td>
<td>Molybdenum</td>
<td>5 ug/L / 200.8</td>
</tr>
<tr>
<td>Thallium</td>
<td>200 ug/L / 200.7</td>
<td>Silver</td>
<td>5 ug/L / 200.8</td>
</tr>
<tr>
<td>Vanadium</td>
<td>10 ug/L / 200.7</td>
<td>Cadmium</td>
<td>0.7 ug/L / 200.8</td>
</tr>
<tr>
<td>Zinc</td>
<td>20 ug/L / 200.7</td>
<td>Tin</td>
<td>30 ug/L / 200.8</td>
</tr>
<tr>
<td>Chromium</td>
<td>5 ug/L / 200.8 (ICPMS)</td>
<td>Antimony</td>
<td>5 ug/L / 200.8</td>
</tr>
<tr>
<td>Nickel</td>
<td>10 ug/L / 200.8</td>
<td>Barium</td>
<td>2 ug/L / 200.8</td>
</tr>
<tr>
<td>Copper</td>
<td>5 ug/L / 200.8</td>
<td>Thallium</td>
<td>1 ug/L / 200.8</td>
</tr>
<tr>
<td>Zinc</td>
<td>10 ug/L / 200.8</td>
<td>Lead</td>
<td>1 ug/L / 200.8</td>
</tr>
<tr>
<td>Arsenic</td>
<td>5 ug/L / 200.8</td>
<td>Uranium</td>
<td>1 ug/L / 200.8</td>
</tr>
<tr>
<td>Selenium</td>
<td>5 ug/L / 200.8</td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

*Note: Reporting limits for sulfate and nitrate/nitrite are subject to change. A sample with a concentration of either analyte greater than certain ranges may need to be diluted to bring the concentration within the analytical ranges of the testing instruments. This dilution results in a proportional increase in the reporting limit.*
<table>
<thead>
<tr>
<th>Analyte</th>
<th>Primary MCL</th>
<th>Secondary MCL</th>
<th>Analyte</th>
<th>Primary MCL</th>
<th>Secondary MCL</th>
</tr>
</thead>
<tbody>
<tr>
<td>Vinyl Chloride</td>
<td>2 ug/L</td>
<td>None</td>
<td>p-Dichlorobenzene</td>
<td>75 ug/L</td>
<td>None</td>
</tr>
<tr>
<td>1,1-Dichloroethylene</td>
<td>7 ug/L</td>
<td>None</td>
<td>o-Dichlorobenzene</td>
<td>600 ug/L</td>
<td>None</td>
</tr>
<tr>
<td>Dichloromethane</td>
<td>5 ug/L</td>
<td>None</td>
<td>1,2,4-Trichlorobenzene</td>
<td>70 ug/L</td>
<td>None</td>
</tr>
<tr>
<td>Trans-1,2-Dichloroethylene</td>
<td>100 ug/L</td>
<td>None</td>
<td>Chloroform (1)</td>
<td>Total 1,2,3,4 = 80 ug/L</td>
<td>None</td>
</tr>
<tr>
<td>Cis-1,2-Dichloroethylene</td>
<td>70 ug/L</td>
<td>None</td>
<td>Bromodichloromethane (2)</td>
<td>Total 1,2,3,4 = 80 ug/L</td>
<td>None</td>
</tr>
<tr>
<td>1,1,1-Trichloroethane</td>
<td>200 ug/L</td>
<td>None</td>
<td>Chlorodibromomethane (3)</td>
<td>Total 1,2,3,4 = 80 ug/L</td>
<td>None</td>
</tr>
<tr>
<td>Carbon Tetrachloride</td>
<td>5 ug/L</td>
<td>None</td>
<td>Bromoform (4)</td>
<td>Total 1,2,3,4 = 80 ug/L</td>
<td>None</td>
</tr>
<tr>
<td>Benzene</td>
<td>5 ug/L</td>
<td>None</td>
<td>Chloride</td>
<td>None</td>
<td>250 mg/L</td>
</tr>
<tr>
<td>1,2-Dichloroethane</td>
<td>5 ug/L</td>
<td>None</td>
<td>Sulfate</td>
<td>None</td>
<td>250 mg/L</td>
</tr>
<tr>
<td>Trichloroethylene</td>
<td>5 ug/L</td>
<td>None</td>
<td>Nitrate/nitrite</td>
<td>10 mg/L</td>
<td>None</td>
</tr>
<tr>
<td>1,2-Dichloro-propane</td>
<td>5 ug/L</td>
<td>None</td>
<td>Fluoride</td>
<td>4 mg/L</td>
<td>2 mg/L</td>
</tr>
<tr>
<td>Toluene</td>
<td>1,000 ug/L</td>
<td>None</td>
<td>Aluminum</td>
<td>None</td>
<td>50 -200 ug/L</td>
</tr>
<tr>
<td>1,1,2-Trichloroethane</td>
<td>5 ug/L</td>
<td>None</td>
<td>Antimony</td>
<td>6 ug/L</td>
<td>None</td>
</tr>
<tr>
<td>Tetrachloroethylene</td>
<td>5 ug/L</td>
<td>None</td>
<td>Arsenic</td>
<td>10 ug/L</td>
<td>None</td>
</tr>
<tr>
<td>Chlorobenzene</td>
<td>100 ug/L</td>
<td>None</td>
<td>Barium</td>
<td>2000 ug/L</td>
<td>None</td>
</tr>
<tr>
<td>Ethylbenzene</td>
<td>700 ug/L</td>
<td>None</td>
<td>Beryllium</td>
<td>4 ug/L</td>
<td>None</td>
</tr>
<tr>
<td>Total Xylenes</td>
<td>10,000 ug/L</td>
<td>None</td>
<td>Cadmium</td>
<td>5 ug/L</td>
<td>None</td>
</tr>
<tr>
<td>Styrene</td>
<td>100 ug/L</td>
<td>None</td>
<td>Chromium</td>
<td>100 ug/L</td>
<td>None</td>
</tr>
</tbody>
</table>
Table A-10, Continued. Analytes, Primary MCLs (A), and Secondary MCLs (B).

<table>
<thead>
<tr>
<th>Analyte</th>
<th>Primary MCL</th>
<th>Secondary MCL</th>
<th>Analyte</th>
<th>Primary MCL</th>
<th>Secondary MCL</th>
</tr>
</thead>
<tbody>
<tr>
<td>Copper</td>
<td>Action level = 1,300 ug/L&lt;sup&gt;(c)&lt;/sup&gt;</td>
<td>1000 ug/L</td>
<td>Selenium</td>
<td>50 ug/L</td>
<td>None</td>
</tr>
<tr>
<td>Iron</td>
<td>None</td>
<td>300 ug/L</td>
<td>Silver</td>
<td>None</td>
<td>100 ug/L</td>
</tr>
<tr>
<td>Lead</td>
<td>Action level = 15 ug/L&lt;sup&gt;(c)&lt;/sup&gt;</td>
<td>None</td>
<td>Thallium</td>
<td>2 ug/L</td>
<td>None</td>
</tr>
<tr>
<td>Manganese</td>
<td>None</td>
<td>50 ug/L</td>
<td>Zinc</td>
<td>None</td>
<td>5,000 ug/L</td>
</tr>
<tr>
<td>Nickel</td>
<td>100 ug/L</td>
<td>None</td>
<td></td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

**Notes:**

(A) Primary MCL = Primary Maximum Contaminant Level, a maximum concentration of a substance (other than lead or copper) allowed in public drinking water due to adverse health effects.

(B) Secondary MCL = Secondary Maximum Contaminant Level, a maximum concentration of a substance suggested for public drinking water due solely to unpleasant characteristics such as bad flavor or stain-causing ability.

(C) Action Level = the maximum concentrations of lead or copper permitted for public drinking water as measured at the user’s end of the system. Water issuing from at least ninety percent of a representative sample of user’s end outlets must contain copper or lead concentrations at or below their respective action levels.

mg/L = milligrams per liter.

ug/L = micrograms per liter.
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